

ATP Measurements in Laboratory Cultures and  
Field Populations of Lake Plankton

By

FRANCIS XAVIER BROWNE

A DISSERTATION PRESENTED TO THE GRADUATE COUNCIL OF  
THE UNIVERSITY OF FLORIDA IN PARTIAL  
FULFILLMENT OF THE REQUIREMENTS FOR THE DEGREE  
OF DOCTOR OF PHILOSOPHY

UNIVERSITY OF FLORIDA  
1971

## ACKNOWLEDGMENTS

The research effort reported in this dissertation has been accomplished with the support of many people. Particular thanks are expressed to Mr. Roger Yorton, Miss Virginia Morgan, and Mrs. Zena Hodor for their helpful assistance.

I would like to thank Dr. W. H. Morgan for his friendship and support during these years. Special thanks go to Jenny Hunt for her assistance in many little things over the years and for one big thing - typing this dissertation.

I wish to thank my Supervisory Committee for their advice and encouragement, but especially for their friendship during this period. Particular gratitude is extended to my principal advisors, Drs. Patrick L. Brezonik and Jackson L. Fox. I also thank the rest of my committee, Drs. Hugh Putnam and Frank Nordlie, for their ideas and insights.

I also wish to thank my parents for all the help and support they gave me throughout the years. For without them, truly, none of this would have been possible.

Finally, I thank my wife, June, for everything: the frustrations, the sacrifices, the added responsibilities she endured. Her encouragement and understanding have made all this possible.

This investigation was made possible through the financial support of the Department of Health, Education, and Welfare, PHS Traineeship Grant No. 5-T01-EC-00035-10.

## TABLE OF CONTENTS

	Page
ACKNOWLEDGMENTS .....	ii
LIST OF TABLES .....	vi
LIST OF FIGURES .....	viii
ABSTRACT .....	x
CHAPTER I. INTRODUCTION .....	1
CHAPTER II. MECHANISM OF ATP METABOLISM IN ALGAE ...	4
CHAPTER III. LITERATURE REVIEW .....	9
Introduction .....	9
Biomass Parameters .....	10
Pigment Concentration .....	10
Particulate Organic Carbon .....	14
Cell Number .....	15
Packed Cell Volume and Dry Weight .....	16
DNA .....	17
Limiting Nutrient Bioassay Methods .....	18
Carbon-14 Uptake .....	19
Enzymatic and Extractive Techniques .....	21
ATP .....	24
CHAPTER IV. EXPERIMENTAL METHODS AND MATERIALS .....	36
Analytical Techniques .....	36
Experimental Materials .....	38
CHAPTER V. METHODS OF ATP ANALYSIS .....	41
CHAPTER VI. EXPERIMENTAL RESULTS .....	55
ATP - Biomass Results .....	55
ATP vs. Chlorophyll $a$ .....	56
ATP vs. Dry Weight .....	68
ATP vs. Cell Number .....	76
ATP vs. Absorbance .....	76
Ecoplankton Study .....	80

	Page
ATP - Response Results .....	80
Light-Dark Study .....	81
pH Studies .....	83
Toxicity Studies .....	88
Nutrient Studies .....	93
CHAPTER VII. DISCUSSION .....	110
CHAPTER VIII. CONCLUSIONS AND RECOMMENDATIONS .....	115
Conclusions .....	115
Recommendations .....	117
BIBLIOGRAPHY .....	118
BIOGRAPHICAL SKETCH .....	130

## LIST OF TABLES

		Page
TABLE 1.	STANDARD FREE ENERGY OF HYDROLYSIS OF COMMON INTRACELLULAR PHOSPHATE COMPOUNDS .....	4
TABLE 2.	COMPARISON OF BIOMASS ESTIMATED BY ATP, CHLOROPHYLL, DNA AND ORGANIC CARBON .....	18
TABLE 3.	ATP AS PERCENT DRY WEIGHT .....	30
TABLE 4.	ATP TO CHLOROPHYLL RATIO FOR VARIOUS ALGAE UNDER DIFFERENT ENVIRONMENTAL CONDITIONS .....	32
TABLE 5.	EFFECT OF INCUBATION ON FIREFLY EXTRACT AND ATP LIGHT EMISSION (6-SEC COUNT) .....	50
TABLE 6.	ATP VS. CHLOROPHYLL <i>A</i> CORRELATION COEFFICIENTS FOR UNIALGAL CULTURES .....	60
TABLE 7.	CHEMICAL AND BIOLOGICAL CHARACTERISTICS OF EXPERIMENTAL LAKES .....	63
TABLE 8.	ATP VS. CHLOROPHYLL <i>A</i> CORRELATION COEFFICIENTS FOR LAKE PHYTOPLANKTON .....	65
TABLE 9.	ATP TO CHLOROPHYLL <i>A</i> RATIOS FOR ALGAL CULTURES AND LAKE PHYTOPLANKTON .....	66
TABLE 10.	ATP VS. DRY WEIGHT CORRELATION COEFFICIENTS FOR UNIALGAL CULTURES .....	71
TABLE 11.	ATP TO DRY WEIGHT CORRELATION COEFFICIENTS FOR LAKE PHYTOPLANKTON .....	74
TABLE 12.	ATP TO DRY WEIGHT RATIOS FOR ALGAL CULTURES AND LAKE PHYTOPLANKTON .....	75
TABLE 13.	COMPARISON OF ATP/DRY WEIGHT RATIO WITH AVERAGE TURBIDITY OF EXPERIMENTAL LAKES ....	75

	Page
TABLE 14. ATP TO ABSORBANCE CORRELATION COEFFICIENTS FOR UNIALGAL CULTURES .....	79
TABLE 15. PERCENT REDUCTION AFTER ONE-HOUR INCUBATION WITH MERCURIC CHLORIDE .....	90
TABLE 16. RESPONSE OF CELLULAR ATP TO ADDITION OF PHOSPHATE .....	99
TABLE 17. RESPONSE OF CELLULAR ATP TO NITRATE ADDITION .....	108
TABLE 18. ATP TO DRY WEIGHT RATIOS OF VARIOUS MARINE ALGAE .....	111
TABLE 19. ATP TO DRY WEIGHT RATIOS OF <i>SELENASTRUM</i> .	112

LIST OF FIGURES

	Page
FIGURE 1. CHEMICAL STRUCTURE OF ATP .....	5
FIGURE 2. CELLULAR CONTENTS OF ATP IN <i>EUGLENA</i> <i>GRACILIS</i> DURING ALTERNATING PERIODS OF LIGHT AND DARK .....	27
FIGURE 3. ATP CONCENTRATION VS. CELLULAR ORGANIC CARBON .....	29
FIGURE 4. LUMINESCENCE DECAY OF ATP FIREFLY LANTERN EXTRACT REACTION WITH TIME .....	48
FIGURE 5. DEPENDENCE OF QUANTITATIVE ACTIVATED SLUDGE ATP EXTRACTION ON TEMPERATURE OF EXTRACTION SOLUTION .....	51
FIGURE 6. STANDARD ATP CURVE .....	53
FIGURE 7. CORRELATION OF ATP WITH <i>SELENASTRUM</i> CONCENTRATION .....	54
FIGURE 8. ATP VS. CHLOROPHYLL A OF <i>SELENASTRUM</i> AND <i>CHLORELLA</i> .....	57
FIGURE 9. ATP VS. CHLOROPHYLL A FOR <i>ANABAENA</i> .....	58
FIGURE 10. ATP VS. CHLOROPHYLL A FOR <i>MICROCYSTIS</i> ....	59
FIGURE 11. ATP VS. CHLOROPHYLL A OF DILUTED <i>MICROCYSTIS</i> .....	62
FIGURE 12. ATP VS. CHLOROPHYLL A FOR LAKE PHYTOPLANKTON .....	64
FIGURE 13. VARIATION OF ATP/CHLOROPHYLL A RATIO WITH CELL AGE .....	69
FIGURE 14. ATP VS. DRY WEIGHT OF UNIALGAL CULTURES ..	70
FIGURE 15. ATP VS. DRY WEIGHT FOR LAKE PHYTOPLANKTON .....	73



	Page
FIGURE 16. ATP VS. CELL COUNT FOR <i>MICROCYSTIS</i> .....	77
FIGURE 17. ATP VS. ABSORBANCE OF UNIALGAL CULTURES ..	78
FIGURE 18. RESPONSE OF ATP TO PERIODS OF LIGHT AND DARK .....	82
FIGURE 19. EFFECT OF pH ON ATP CONTENT OF <i>SELENASTRUM</i> CELLS .....	84
FIGURE 20. EFFECT OF pH ON ATP CONTENT OF ANDERSON- CUE LAKE WATER .....	86
FIGURE 21. EFFECT OF pH ON ATP CONTENT OF BIVENS ARM LAKE WATER .....	87
FIGURE 22. ATP POOL RESPONSE TO INCUBATION WITH VARIOUS CONCENTRATIONS OF MERCURY .....	89
FIGURE 23. RESPONSE OF MERCURY-POISONED <i>CHLORELLA</i> AFTER TRANSFER TO FRESH MEDIUM .....	92
FIGURE 24. RESPONSE OF BIOMASS PARAMETERS IN BIVENS ARM TO ADDITION OF COPPER .....	94
FIGURE 25. ATP RESPONSE IN <i>ANABAENA</i> TO ADDITION OF CARBON DIOXIDE .....	96
FIGURE 26. ATP RESPONSE OF BIVENS ARM TO ADDITION OF CARBON DIOXIDE .....	98
FIGURE 27. RESPONSE OF ATP IN <i>ANABAENA</i> TO ADDITION OF PHOSPHORUS .....	101
FIGURE 28. RESPONSE OF ATP IN <i>MICROCYSTIS</i> TO ADDITION OF PHOSPHORUS .....	102
FIGURE 29. RESPONSE OF ATP IN <i>SELENASTRUM</i> TO ADDITION OF PHOSPHORUS .....	103
FIGURE 30. RESPONSE OF ATP IN <i>CHLORELLA</i> TO ADDITION OF PHOSPHORUS .....	104
FIGURE 31. RESPONSE OF ATP IN <i>ANABAENA</i> TO ADDITION OF NITROGEN .....	106
FIGURE 32. RESPONSE OF ATP IN <i>SELENASTRUM</i> TO ADDITION OF NITROGEN .....	107

Abstract of Dissertation Presented to the Graduate Council  
of the University of Florida in Partial Fulfillment of the  
Requirements for the Degree of Doctor of Philosophy

ATP MEASUREMENTS IN LABORATORY CULTURES AND  
FIELD POPULATIONS OF LAKE PLANKTON

By

Francis Xavier Browne

December, 1971

Chairman: Patrick L. Brezonik, Ph.D.

Co-Chairman: Jackson L. Fox, Ph.D.

Major Department: Environmental Engineering

ATP measurements in laboratory unialgal cultures and lake plankton were made to investigate the use of ATP as a biomass and activity parameter. No account was taken of the bacterial population in the unialgal cultures or the lake plankton. Although the algal cultures were not axenic, efforts were made to minimize bacterial growth. ATP was analysed using the luciferin-luciferase firefly reaction. The ATP analysis proved to be sensitive and reliable for quantitative determination of cellular ATP concentrations.

Good correlation was observed between ATP and chlorophyll  $a$ , dry weight, and cell number. ATP to chlorophyll  $a$  ratios were relatively constant, ranging from 0.09 to 0.35. These values agree with those reported by other investigators. ATP to dry weight ratios were also relatively constant. The presence of detrital material in laboratory and field samples reduced the significance of any particular ATP to dry weight ratio.

The ATP concentration was observed to be relatively constant under both light and dark conditions. Maximum cellular ATP concentrations occurred at a pH level at or in the range of the normal pH of the organisms tested.

Rapid decreases in ATP concentration occurred immediately after addition of toxic substances. Results of one experiment indicate that toxic substances, at least in the first few hours, cause a reduction in ATP per cell rather than a reduction in viable biomass.

Additions of nutrients to nutrient-deficient cultures resulted in rapid increases in ATP concentration.

The advantages of ATP over existing biomass parameters are manifold. The results of this research indicate that ATP analysis is applicable as a biomass parameter for aquatic systems. It is also a sensitive activity parameter which can be applied to toxicity bioassays. These results also indicate that ATP could be used as a qualitative activity parameter in limiting nutrient bioassays.

## CHAPTER I. INTRODUCTION

Pollution of surface waters is well recognized as a serious problem in the United States and other highly developed countries. The effects of pollution range from the death of aquatic organisms as a result of oxygen depletion from organic discharges or poisoning from toxic industrial waters to over population with nuisance organisms as a result of enrichment from nutrient discharges. To properly measure and evaluate the effects of various pollutants on a particular water, reliable measures of biomass and activity are needed. The biomass parameters most often used in lake and stream studies are chlorophyll concentration, suspended solids, or plankton count. The application, theory and problems associated with these parameters are discussed in Chapter III, but it suffices to say here that better measures of phytoplankton biomass and activity are needed.

The deficiencies of present methodologies are evidenced by the many conflicting reports in the literature concerning limiting nutrients and the effects of municipal and industrial discharges on a body of water. In 1969 a Provisional Algal Assay Procedure (Joint Industry-Government Task Force on Eutrophication 1969) was developed to

standardize existing methods of investigating phytoplankton responses in natural waters. However, this procedure has not been wholly successful probably because it utilizes inadequate biomass parameters.

An adequate biomass parameter must have a relatively constant cellular concentration under most environmental conditions, and it must not be associated with non-living material. Cellular biochemistry offers a number of possibilities for measurement of activity and biomass. One biochemical parameter that seems to offer an appropriate measure of biomass and metabolic activity is adenosine triphosphate (ATP). The overall objective of this study was to investigate the use of adenosine triphosphate (ATP) as a measure of phytoplankton biomass and metabolic activity. Biomass parameter evaluation was divided into two phases. First, the ATP content of batch unialgal cultures was measured and correlated with traditional biomass parameters. In the second phase, the ATP content of natural lake phytoplankton was measured and correlated with biomass parameters. ATP as a measure of metabolic activity was evaluated by observing the ATP changes in laboratory algal cultures and lake phytoplankton which were subjected to varying environmental conditions.

Correlation of cellular ATP concentration with current biomass parameters would indicate the validity of using ATP to measure phytoplankton biomass. The lack of response of

cellular ATP to various environmental conditions would indicate the stability of cellular ATP to minor environmental changes. The response of cellular ATP to additions of nutrients and toxic substances would indicate its usefulness as a rapid bioassay parameter.

## CHAPTER II. MECHANISM OF ATP METABOLISM IN ALGAE

ATP has been called the "energy currency" of living cells. It occurs in all living cells. ATP is a nucleotide containing adenine, a 6-amino derivative of purine, D-ribose, a 5-carbon sugar, and three phosphate groups as shown in Figure 1. The ATP molecule within the cell is highly charged with negative charges concentrated around the polyphosphate structure. Most of the ATP in cells is present as a  $Mg^{+2}$  complex.

It is common practice to refer to ATP as a "high energy" phosphate compound. However, as shown in Table 1, ATP has an intermediate energy value when compared with other phosphate compounds. It is this intermediate position which makes ATP so important.

TABLE 1

### STANDARD FREE ENERGY OF HYDROLYSIS OF COMMON INTRACELLULAR PHOSPHATE COMPOUNDS

	$\Delta G^\circ$ Kcal/Mole
Phosphoenolpyruvate	-14.80
1,3-Diphosphoglycerate	-11.80
Phosphocreatine	-10.30
Acetyl phosphate	-10.10
Phosphoarginine	- 7.70
ATP	- 7.30
Glucose 1-phosphate	- 5.00
Fructose 6-phosphate	- 3.80
Glucose 6-phosphate	- 3.30
Glycerol 1-phosphate	- 2.20

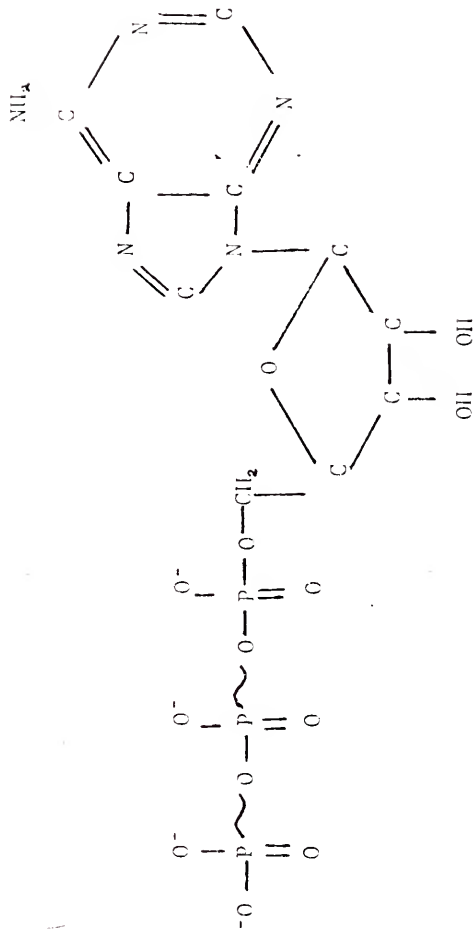


FIGURE 1. CHEMICAL STRUCTURE OF ATP



The function of the ATP-ADP (adenosine diphosphate) system is to act as an intermediate between high-energy phosphate compounds and low-energy phosphate compounds. One set of enzymes helps transfer phosphate from high-energy compounds to ADP, forming ATP; while another set of enzymes help transfer the terminal phosphate group of ATP to low-energy phosphate acceptors. No enzymes exist that can transfer phosphate groups directly from high-energy donors to low-energy acceptors. Thus, all high-energy phosphate transfers must use the ATP-ADP system.

Inside living cells there is usually very little ATP since it is continually being used up and remade. The energy in the energy-rich bonds of ATP is used to do cellular work. It is used to synthesize large molecules from simple subunits, to transport substances in and out of the cell, and to perform mechanical work.

In algae, ATP is formed from two basic processes: photosynthesis and respiration. Eucaryotic algal cells contain chloroplasts, where photosynthesis occurs, and mitochondria, where respiration occurs. Procaryotic cells, like the blue-green algae, do not have chloroplasts and mitochondria. Although eucaryotic and procaryotic cells differ in structure, the chemical reactions occurring in photosynthesis and respiration are similar.

Photosynthesis consists of both light and dark reactions; the light reactions drive the entire process. In the light reactions, water is split, ATP is formed, and

oxygen is liberated. The dark reactions use carbon dioxide to form simple sugars, some of which are converted into glucose as the product of photosynthesis. In photosynthesis, chlorophyll molecules absorb light energy and give off high-energy electrons. It is believed that molecules of ferredoxin, a protein which contains iron, pick up the high-energy electrons. The energy gained by ferredoxin is released in a series of electron transfers. This released energy is used to form ATP. The electrons, after losing their energy, return to the chlorophyll molecule again. This process is called cyclic photophosphorylation. Another process called non-cyclic photophosphorylation also occurs and is similar to the cyclic process except the original electrons do not return to the chlorophyll molecule. Instead, the electrons are transferred to NADP (nicotinamide adenine dinucleotide phosphate) which subsequently is reduced. Electrons from the hydroxyl groups of water are transferred to the chlorophyll molecules, filling the electron gaps. Thus the primary purpose of photosynthesis is to form ATP and reduced NADP, with oxygen evolved as a by product. Under normal conditions one would expect non-cyclic photophosphorylation to occur since this process forms ATP and reducing power (in the form of reduced NADP). Both ATP and the reducing power are used in the dark reactions to form glucose from  $\text{CO}_2$ . The role of cyclic photophosphorylation is not completely understood. Some believe it is a shunt mechanism used to

form ATP without formation of reduced NADP. It appears that light-induced cyclic electron flow is used to generate ATP at whatever rate is required by the cell, without the necessity of generating reduced NADP or of evolving molecular oxygen.

ATP production from respiration is basically the same for algae and other microorganisms. Catabolic processes which terminate in the Krebs Cycle product ATP. For example, one molecule of glucose is transformed into two molecules of pyruvic acid, forming two molecules of ATP. These two molecules of pyruvic acid are broken down into six molecules of carbon dioxide and water, liberating 36 molecules of ATP. Thus, a net ATP production of 38 molecules of ATP results from the oxidative breakdown of one glucose molecule.

## CHAPTER III. LITERATURE REVIEW

### Introduction

Vollenweider (1970) defines eutrophication as anything that accelerates the nutrient loading, increases the nutrient level and directly increases water productivity. He explains that a mere increase in the nutrient level is not important if it does not produce unpleasant or undesirable effects on the metabolism of the water. Vollenweider's definition emphasizes the two primary constituents of eutrophication - the addition of nutrients to a water and the concomitant response of the phytoplankton. Measuring the standing phytoplankton biomass and its response to nutrient additions is becoming increasingly important; yet, there is little consensus as to which parameter or method is best to use. A variety of methods is presently used to measure biomass and its response to nutrient additions, including pigment concentration, particulate organic carbon concentration, cell number or areal units, packed cell volume and dry weight, and DNA concentration. While commonly used each suffers from certain deficiencies. The following sections will briefly describe the merits and drawbacks of each.

## Biomass Parameters

### Pigment Concentration

The algal pigments consist primarily of the chlorophylls, the carotenoids, and the phycobilins. Chlorophylls are lipo-soluble molecules characterized by strong absorption of red (650 - 680 m $\mu$ ) and blue light (400 - 450 m $\mu$ ) and by their red fluorescence in organic solvents. Chlorophylls are divided into three groups: chlorophyll *a* which is present in all photosynthesizing cells; chlorophyll *b*, an oxidized derivative of chlorophyll *a*, which is found in green algae; and chlorophyll *c* which is found in brown algae and diatoms. Carotenoids are C<sub>40</sub> hydrocarbons with numerous conjugated double bonds. There are two major groups of carotenoids: the carotenes, which are pure hydrocarbons, and the xanthophylls, which contain from 1 to 8 hydroxyl groups making them more polar. The carotenoids serve as accessory pigments, extending the range of visible light useful in photosynthesis. They are usually yellow or orange and absorb light in the blue-violet region of the spectrum. All photosynthesizing cells contain carotenoids. Odum and Nixon (1970) have studied the role of carotenoids in photorespiration. Odum *et al* (In Press) note that the carotenoid-chlorophyll ratio may help measure the relative use of solar energy in photosynthesis and photorespiration. The third major class of pigments are the phycobilins which are related structurally to the

chlorophylls. They consist of an open conjugated system of four pyrrol rings, the fundamental structure of the bilin pigments, so called because they were discovered in bile. The phycobilins (algal bilins) are water-soluble pigments consisting of the phycoerythrins and the phycocyanins. Both pigments are found in the red marine algae (Rhodophyceae) and the primitive blue-green algae (Cyanophyceae). Phycoerythrin, present primarily in red marine algae, absorbs light in the middle of the visible spectrum. Phycocyanin, present primarily in blue-green algae, absorbs red light at about 650 m $\mu$ .

Chlorophyll is the major pigment of algae and occurs in all species. Because of its universal distribution, chlorophyll  $\alpha$  is often used to estimate phytoplankton biomass. Weber and McFarland (1969) and Keup and Stewart (1966) used chlorophyll  $\alpha$  as an estimate of the phytoplankton standing crop and found it to correlate well with carbon, nitrogen, and phosphorous content.

One problem with using chlorophyll as a measure of photoplankton biomass is that the chlorophyll concentration varies with environmental conditions. Spoehr and Milner (1949) measured extremes of 0.1 percent and 6 percent for chlorophyll content as percent of dry weight. Usually the chlorophyll content is about 0.5 to 1.5 percent of the dry weight (Round 1967). Studies on algae cultured under controlled laboratory conditions demonstrate that mineral nutrition, light intensity, and cell age affect the cellular

chlorophyll concentration. Emerson (1929) found a direct correlation between photosynthetic activity and chlorophyll content in *Chlorella pyrenoidosa* and concluded that a reduced chlorophyll content was the only factor responsible for the reduced photosynthesis of iron-deficient cells. Sargent (1940) noted that chlorophyll constituted 6.6 percent of the dry weight of "shade-grown" *Chlorella* cells but only 3.3 percent of "sun-grown" cells. Myers (1946) found that chlorophyll per milliliter of packed cells was inversely related to light intensity, but he also found a direct relationship between light intensity and cell size. Fogg (1965) observed that the cellular chlorophyll concentration decreased with cell age and nutrient deficiency. Ryther and Yentsch (1957) found that marine phytoplankton at light saturation has a reasonably constant assimilation ratio of 3.7 grams of carbon assimilated per hour per gram of chlorophyll. Calculated production rates based on this ratio and on chlorophyll-light measurements were similar to those obtained by simultaneous use of the light-and-dark bottle oxygen method. Bain (1968) used a similar method to predict primary production in San Francisco Bay. Some controversy exists as to whether chlorophyll *a* can best be used to measure biomass or primary production (Odum *et al* 1958).

The analytical technique for extracting and calculating chlorophyll concentration suffers from many weaknesses. Strickland and Parsons (1968) note that the only

rapid chemical method known for estimating living plant matter in the particulate organic matter of sea water is to determine the characteristic pigments. However, they also note that the amount of organic material associated with a given plant pigment is variable, depending on the class of algae and the nutritional state. A factor ranging from 25 to 100 is used to convert chlorophyll *a* concentration to total plant carbon. According to Strickland and Parsons (1968) extraction with 90 percent acetone gives results undoubtedly low in many instances because of the presence of plant cells that are not fully extracted. These authors state that some species may retain 50 percent or more of the pigments in their cells. Bogarad (1962) notes that it is generally more difficult to extract chlorophylls from algae than from higher plants; he adds that extraction with hot or cold methanol is usually more effective than acetone. Another weakness of the extraction process is that inactive chlorophyll and degradation products may be determined along with the active chlorophyll of the living phytoplanktons. Glooschenko and Moore (1971) analyzed Lake Ontario water samples for chlorophyll *a* and its degradation products, pheophorbide *a* and pheophytin *a*. These authors found the degradation products made up less than 20 percent of total chlorophyll *a* until the decline of the spring growth, after which degradation products accounted for up to 100 percent measured chlorophyll *a* in some regions of the lake. These studies indicate the necessity of correcting chlorophyll *a* data for degradation products.



Particulate Organic Carbon

Particulate organic carbon is often used as a measure of phytoplankton biomass. Menzel and Ryther (1964) measured the composition of particulate matter in the western North Atlantic to determine the relationship between carbon, nitrogen, phosphorus, and chlorophyll. Regression of phosphorus or chlorophyll vs. nitrogen or carbon, when extrapolated back toward the origin, indicated appreciable amounts of nitrogen and carbon in the absence of phosphorus and chlorophyll. In contrast, the regression of phosphorus vs. chlorophyll had its intercept at the origin. The authors concluded that chlorophyll and phosphorus are decomposed or mineralized at about the same rate, while carbon and nitrogen are more refractory. Menzel and Goering (1966) studied the decomposition of particular matter from surface and deep Atlantic waters. Changes in carbon were related to the initial biomass of phytoplankton. They found that the living carbon represents a variable fraction of the total organic particulate matter present in surface waters. Superimposed on this is detrital carbon which is refractory to decomposition. Parsons and Strickland (1962), Holm-Hansen (1969) and others have also found variable amounts of refractory particulate organic carbon in phytoplankton samples. Because of this refractory nature of carbon its use as an estimate of biomass is questionable.

Cell Number

Early work on algal physiology utilized the cell number, determined by counting cells with a hemocytometer, whipple disc, or a Segwick-Rafter Cell (Myers 1962). With unialgal cultures the cell count can be relatively meaningful because of the approximate uniformity of the cell size and volume. Cell counts on a heterogeneous culture of algae are tedious and open to misinterpretations. Oswald and Gaonkar (1969) in a review of the Provisional Algae Assay Procedure (Joint Industry-Government Task Force on Eutrophication 1969) note the difficulty in counting cells in mixed populations: taking the size of *Chlorella pyrenoidosa* as unity, the size of *Scenedesmus obliquus* is 1.2, *Chlamydomonas* 6.9, *Euglena gracilis* 54, *Euglena viridis* 84, and a *Phaeus* sp. 71. Besides this variability among species, cell size differs with both age and nutritional state of each species. *Euglena* cells tend to become larger with age while *Chlorella* and *Scenedesmus* cells become smaller. Nutrient starved *Euglena* usually decrease in size while nutrient starved *Chlorella* increase in size. There is also a difficulty in counting filamentous and agglomerated cells like *Anabaena* and *Microcystis*. Oswald and Gaonkar also note that many algae are photo regulated. *Chlorella*, for example, accumulates nutrients during the day and divides during the night; thus, unless cells are counted at the same time each day, the number of cells could vary significantly.

Aside from the problem of size, there are difficulties with counting technique. Hee (1954) studied variability in cell counts using the hemocytometer, a popular and simple cell counting device. He found that for *Euglena* the variation in cell counts exceeds 30 percent when the number of cells counted in the hemocytometer grid is less than 30, but that clumping occurs when the number of cells exceed 100. For *Scenedesmus* he found that a minimum of about 15 percent variation occurs when the number of cells counted in the five central squares exceeds 80. The variation is a minimum for *Chlorella* when the cells in the five squares exceed 100.

Finally, a problem exists with respect to count variations among technicians (Oswald and Gaonkar 1969). Although this problem occurs with all laboratory techniques, it is particularly acute for cell counting since personal judgement and knowledge of algae identification play a major role.

#### Packed Cell Volume and Dry Weight

Oswald and Gaonkar (1969) recommended the volumetric technique, commonly known as the packed-cell volume, be added to the Provisional Algal Assay Procedure as a method for evaluating algal growth. As Oswald (1967) noted, the volumetric (packed-cell volume) technique is a non-destructive procedure in which the sample need not be discarded after analysis and can be used for further analysis.

According to Oswald, the packed volume usually contains about 14 percent dry weight of algae and the approximate algal content in dry weight is 1400 times the packed cell volume plus or minus 400. Myers (1962) observed that sufficient centrifugal force and time must be used to obtain a constant and minimum packed cell volume. A major objection to the packed cell volume as a measure of biomass is that it measures detrital as well as living matter. Lee *et al* (1971) found a significant amount of detrital material in unialgal cultures of *Selenastrum capricornutum* after only four days of growth in batch systems. Thus, quantitative interpretation of packed cell volume results can be misleading.

### DNA

Deoxyribonucleic acid (DNA) consists of two polynucleotide chains twisted upon each other to form a double helix; it is the genetic substance which contains the hereditary information of the cell. The use of DNA as a measure of phytoplankton biomass has been investigated by Holm-Hansen and colleagues (1968). After measuring concentrations in the Atlantic and Pacific Oceans and comparing them with measurements of chlorophyll and organic carbon, the authors concluded that there is either a considerable quantity of living material that is high in DNA or that DNA is associated with particulate, non-living material. Holm-Hansen (1969) measured ATP, chlorophyll, DNA, and organic

carbon off the coast of California; he converted ATP, chlorophyll, and DNA concentrations to organic carbon using conversion factors of 250, 50, and 100 respectively as shown in Table 2.

TABLE 2  
COMPARISON OF BIOMASS ESTIMATED BY ATP, CHLOROPHYLL, DNA  
AND ORGANIC CARBON

Biomass ( $\mu\text{g c/liter}$ )	Depth (m)			
	0	50	100	200
Direct Examination	13	25	5.8	1.1
ATP	24	22	5.6	1.9
Chlorophyll	14	22	5.0	0.2
DNA	250	200	95	50

The biomass estimates as determined from DNA measurements are, according to Holm-Hansen, "impossibly high." He concluded that DNA is an excellent biomass parameter for laboratory cultures but not for work in the natural environment.

None of the methods discussed in this section are very good estimates of biomass. With biomass, nothing is constant or absolute; all parameters vary and this variation makes it difficult to say which is the best measure of biomass.

#### Limiting Nutrient Bioassay Methods

Many methods are used to determine which physical or chemical conditions limit phytoplankton growth. One general approach is to monitor the phytoplankton response over a

period of time after enriching the sample with some nutrient. Although popular, this method depends on the response parameter measured and may give misleading results if another nutrient or condition becomes limiting. Response parameters include the various biomass measures and  $^{14}\text{C}$  assimilation rates. The former are usually insensitive to short term changes while insufficient time for response to enrichment may give inaccurate results with the latter. Other nutrient bioassay procedures are based on analysis of planktonic constituents (e.g., adaptive enzymes, percent nitrogen or phosphorus) which vary in response to nutritional conditions.

#### Carbon-14 Uptake

Carbon assimilation rates, as measured by the fixation of carbon-14, are often used as a measure of algal growth in bioassays (Goldman 1960, 1961, 1962, 1964, 1965, 1967; Goldman and Carter 1965; McAllister *et al* 1964; Menzel and Ryther 1960, 1961; Putnam 1966). Primary production measured by the carbon-14 method is a function of the environmental conditions prevailing at the time of incubation. Light intensity, for instance, is an important factor in the amount of carbon fixed. After incubation formaldehyde is usually added to kill the organisms, but Strickland and Parsons (1968) note that even small concentrations of formaldehyde may affect the excretion or loss of organic material from delicate algae. *In situ* measurements of productivity require two to three hours of incubation, thus

limiting the number of sampling stations that can be tested in one day. On the other hand, laboratory incubation using a constant light intensity gives more precise but inaccurate results. Oswald and Gaonkar (1969) state that the technique of radiocarbon measurements as set forth in PAAP seems needlessly complex, delicate, and subject to error in the hands of inexperienced personnel.

Goldman (1960, 1962, 1963) used carbon-14 uptake for bioassays of limiting nutrients. Carbon-14 was added to a phytoplankton culture to which nutrients were added. Periodically, subsamples were removed and the uptake of carbon-14 measured. This method measured the total carbon taken up from the beginning of the experiment. An alternate method consisted of adding nutrients to an unlabeled phytoplankton culture and periodically removing a subsample, adding carbon-14 to it, incubating for three to six hours, and then measuring the amount of carbon taken up during the incubation period. Interpretation of the results of these methods can be misleading due to the inherent variation in measuring carbon-14 uptake. Since net productivity is not being measured by this method, it is questionable what exactly the results mean; is an increase in biomass also occurring?

Many basic questions concerning the carbon-14 method must be answered. Is there any luxury uptake of carbon; how much fixed  $^{14}\text{C}$  is excreted, how much is internally recycled? What is the difference between  $^{12}\text{C}$  and  $^{14}\text{C}$

uptake? Strickland and Parsons (1968) use the factor of 1.05 to account for this difference, but note that this value is uncertain.

### Enzymatic and Extractive Techniques

Enzymatic and extractive methods of analysis have been used to study limiting nutrient concentrations in algal cultures (Fitzgerald 1966, 1969; Shapiro and Ribeiro 1965). Surplus phosphorus, the internal concentration in excess of the amount needed for maximum growth, can be extracted from algae by boiling the algal sample in water for 60 minutes and measuring the orthophosphate in the extract. Fitzgerald (1969) used this procedure and demonstrated that algae limited by phosphorus contain little or no extractable phosphate while algae grown with surplus phosphorus released more than 0.08 mg  $PO_4$ -P/100 mg algae. Gerloff and Skoog (1954, 1957) used a tissue analysis method to evaluate nutrient availability in Wisconsin lakes for the growth of *Microcystis aeruginosa*. Tissue analysis was also used to measure the nutrient availability for the growth of angiosperm aquatic plants (Gerloff and Kromholz 1966). The tissue analysis method requires the establishment of a critical level for each element. The critical level is the minimum tissue content in a particular species that is required for maximum growth. Tissue contents below the critical concentration are associated with deficiencies of that element resulting in less than maximum yields. This method requires the



determination of the critical level for each species encountered; the maximum growth rate for each species is a function of many conditions besides nutrient levels (light, temperature, and pH for instance), and may be difficult to evaluate. In mixed phytoplankton populations the complexity of critical levels tend to make the method impractical. Gerloff and Skoog (1954, 1957), using tissue analysis on *Microcystis aeruginosa* in Wisconsin lakes, concluded that nitrogen was more likely to become growth limiting than was phosphorus. However, Gerloff and Krombholz (1966), using tissue analysis on angiosperm aquatic plants in the same Wisconsin lakes, concluded that phosphorus was more likely to limit higher aquatic plant growth than nitrogen. These results demonstrate the species-dependency of the tissue analysis technique, and highlight the qualitative and relative nature of the method.

Alkaline phosphatase activity is another measure of algal phosphorus nutrition. When algae are phosphorus limited, the alkaline phosphatase activity per unit weight is as much as 25 times that of algae grown with surplus phosphate (Fitzgerald 1966, 1969). Fitzgerald cautions that the effect of the local environment such as recent rains or unusual circulation patterns in lakes cause changes in the distribution of algae and must be considered when interpreting nutritional data. He also notes that results vary according to the species of algae under consideration.

Fitzgerald (1968) found that the rate of  $\text{NH}_4\text{-N}$  absorption by algae in the dark is 4-5 times greater for algae which are nitrogen limited compared to plants with available nitrogen. The comparative rate of ammonium nitrogen absorption in the dark of algae containing surplus nitrogen versus algae limited by available nitrogen was measured. The test consisted of placing 5-20 mg (dry weight) of algae, washed in nitrogen-free medium, into 10-30 ml of Gorham's medium (minus N) and adding 0.1 mg ammonium nitrogen. After one hour incubation at 25° C in the dark the ammonium nitrogen content of the supernatant was compared to controls not containing algae. The procedure is based on the fact that nitrogen-starved cells can assimilate ammonium-nitrogen in the dark while normal cells require light and carbon dioxide (Wetherell 1958). It is believed that nitrogen-starved cells have a carbohydrate reserve which is lacking in normal cells and nitrogen-deficient cells assimilate ammonium-nitrogen until their carbohydrate reserves are exhausted (Syrett 1962).

All of the above techniques whether they be extractive or enzymatic are best interpreted when applied to unialgal test cultures rather than natural waters. These methods give results which are relative and qualitative, and standardization is difficult, that is, a given number by itself means little and interpreting what it means is no easy task.

ATP

In the past, most studies on the ATP pool in microorganisms were confined to cultures of bacteria grown on synthetic substrates. By using pure cultures of bacteria in a defined stage of growth, biochemists have been able to study the role of ATP in the bioenergetics of the cell. Recently, however, some researchers have proposed that ATP could be used to measure microbial biomass. Determination of microbial biomass by measurement of ATP depends upon the assumptions that ATP is not associated with non-living particulate material and that the ratio of ATP to cell carbon is fairly constant (Holm-Hansen and Booth 1966). It is also important that the cellular ATP pool does not vary substantially under different environmental conditions.

Forrest and Walker (1965) observed that the ATP pool in starved cultures of *Streptococcus faecalis* remained constant for almost three hours under endogenous conditions. They concluded that an energy balance kept the ATP pool constant until all the stored substrate was utilized. They also found that the length of time the ATP pool remained constant during endogenous conditions was proportional to the initial substrate concentration.

Bauchop and Elsdén (1960), working with three species of bacteria, showed that under anaerobic conditions the amount of ATP synthesized was proportional to new cell yield. Elsdén (1963) reported that the "ATP growth

coefficient," that is, the grams of new cells produced per mole of ATP synthesized, was approximately constant for all the organisms studied. D'Eustachio *et al* (1968) reported that cell counts based upon ATP concentration were linearly correlated to standard plate count. D'Eustachio and Levin (1967), studying the ATP pool in three aerobic bacterial species during lag, exponential and stationary growth phases, found that *E coli* had a relatively constant level of ATP throughout all growth phases. *Pseudomonas fluorescens* and *Bacillus subtilis* were also fairly constant except for a small increase in ATP pool during exponential growth.

Holm-Hansen and his co-workers have used ATP to measure phytoplankton in the ocean (Holm-Hansen and Booth 1966; Hamilton and Holm-Hansen 1967; Holm-Hansen, Sutcliffe and Sharp 1968; Holm-Hansen 1969, 1970). Experiments have shown that ATP is not associated with non-living material (Holm-Hansen and Booth 1966). These experiments included killing of various algae and bacteria with heat, repeated freezing, or cyanide. The measured residual ATP was negligible. The ATP content in three cultures of bacteria studied averaged between 0.1 and 0.2 percent of the dry weight while the content in eight species of algae ranged from 0.003 to 0.016 percent of the dry weight. (Holm-Hansen and Booth 1966.) These were maximum variations in ATP over a wide variety of growth conditions and stages in batch cultures.

Hemilton and Holm-Hansen (1967) determined the ATP content of seven marine bacterial isolates cultured in both batch and chemostat conditions. The range of ATP in the chemostat grown cells was  $0.5$  to  $6.5 \times 10^{-9}$   $\mu\text{g}$  ATP/cell, or  $0.3$  to  $1.1$  percent of the cell carbon. Senescent cells in batch cultures and starved cells in general had an ATP content about one-fifth that of exponentially growing cells. Averaging a representative number of observations from the chemostat and batch grown cells, the authors calculated the average ATP of these bacteria to be  $1.5 \times 10^{-9}$   $\mu\text{g}$  ATP/cell. On a per unit cellular carbon basis the ATP was calculated to be  $0.4$  percent of the cell carbon.

To determine whether the cellular ATP of algae changed with light and dark periods, Holm-Hansen (1970) studied the response of algal cells to periods of light and dark. As shown in Figure 2, an initial decrease in ATP pool size occurred in the dark followed by a gradual increase to approximately the initial ATP pool level. His test, however, only covered a time scale of 22 minutes and doesn't conclusively demonstrate the constant nature of cellular ATP under light and dark conditions.

Holm-Hansen (1969) measured the total particulate carbon, nitrogen, ATP, DNA, and chlorophyll in profiles to 600 meters and 1000 meters off the coast of southern California. As shown in Table 2, the biomass estimates based on ATP measurements are in good agreement with those based on chlorophyll data and direct microscopic measurements.

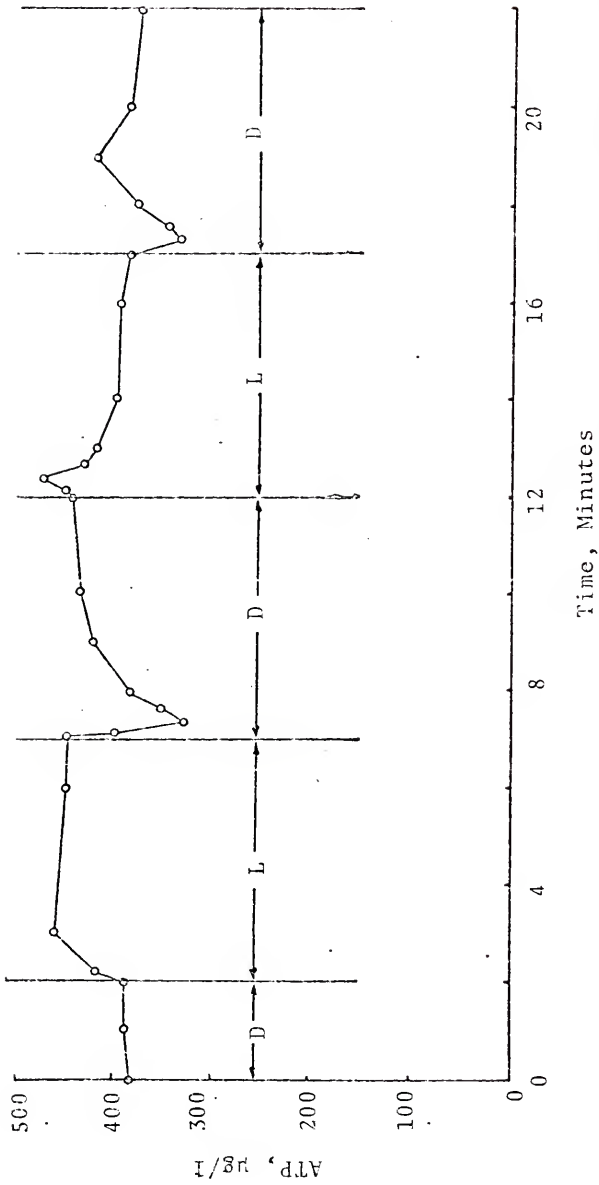


FIGURE 2. CELLULAR CONTENTS OF ATP IN *EUGLENA GRACILIS* DURING ALTERNATING PERIODS OF LIGHT AND DARK (Holm-Hansen 1970)

Holm-Hansen concluded that there is a considerable uniformity in algal ATP concentration over the size range from 1 pg c/cell to 215,000 pg c/cell. Further the average algal ATP value (0.35 percent of the organic carbon) is close to those concentrations reported for bacteria.

Holm-Hansen (1970) investigated the cellular ATP content in 30 different algal cultures under different environmental conditions. As shown in Figure 3, the average concentration of ATP as a percent of the cellular organic carbon is 0.35 percent during exponential growth. Extreme nitrogen deficiency in cultures dropped the ATP level to 35 percent of that found during exponential growth for *Skeletonema costatum*, to 46 percent for *Monochrysis lutheri*, and to 14 percent for *Dunaliella tertiolecta*. Phosphate deficient cultures of *M. lutheri* showed that ATP dropped to about 0.05 percent of the carbon content, but that it increased to 0.15 percent one day after phosphate was added. Holm-Hansen maintains that although ATP pool size varies significantly under extreme nutrient deficient conditions, these conditions would rarely be found in nature. He notes that the apparent change in ATP concentration could result from a significant amount of detrital carbon in his cultures, giving the appearance of a drop in cellular ATP.

The ATP content of microorganisms calculated as percent of dry weight seems relatively constant. Table 3 shows some of the values reported for bacteria and algae. For natural lake and ocean samples of ATP per dry weight

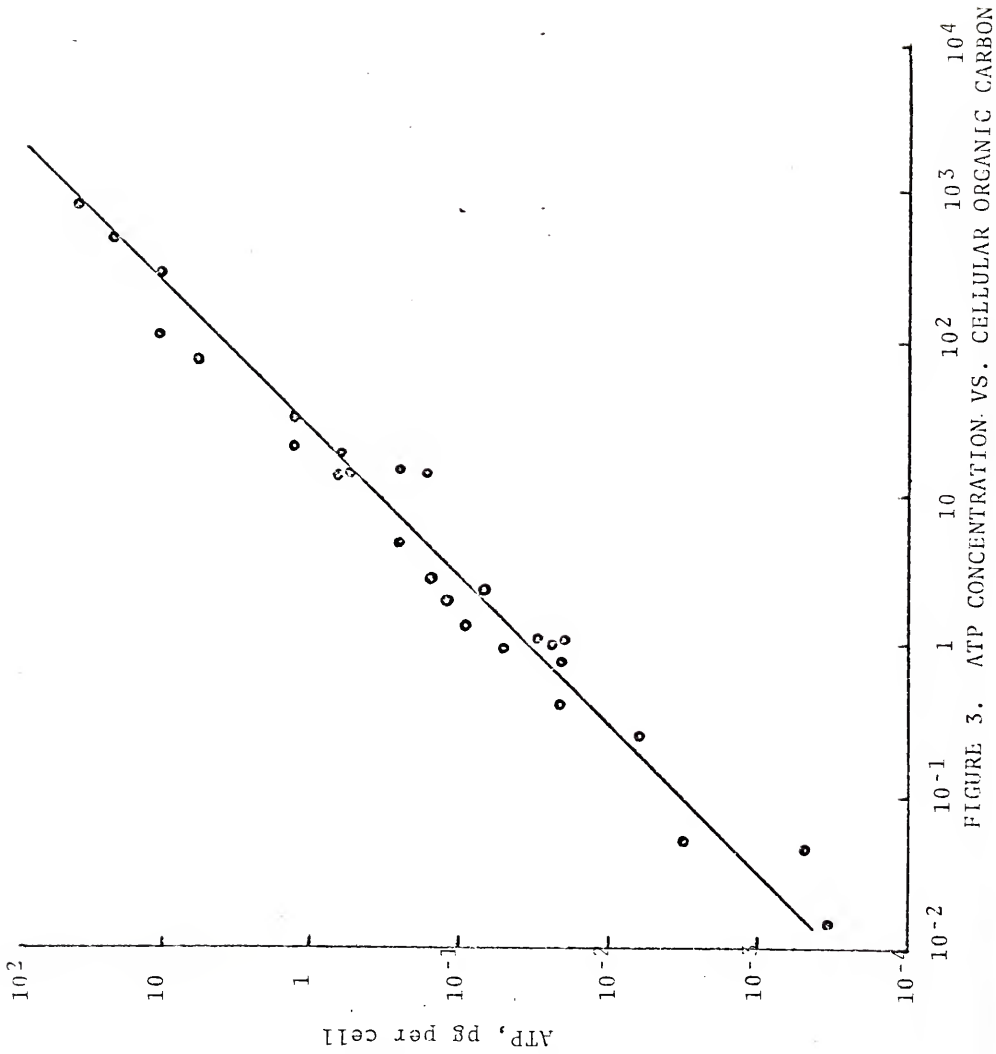


FIGURE 3. ATP CONCENTRATION VS. CELLULAR ORGANIC CARBON



TABLE 5  
ATP AS PERCENT DRY WEIGHT

Organism	$\mu\text{g ATP/mg Dry Weight}$	Reference
<i>Chlorella unalgaris</i> (algae)	2.0	Syrett (1958)
<i>S. cerevisial</i> (bacteria)	0.4	Grylls (1961)
<i>E. coli</i> (bacteria)	3.0	Lehninger (1965)
<i>Strept. faecalis</i> (bacteria)	2.0 - 12.0	Forest (1965)
<i>Salenastrum capricornutum</i> (algae)		Lee (1971)
- phosphorus = 0.62 mg/l	3.1	
- phosphorus = 0.062 mg/l	1.4	
- phosphorus = 3.1 mg/l	3.4	

value may be meaningless because of the presence of detrital material. Holm-Hansen (private communication) does not believe it is feasible to report ATP/dry weight ratios and reports most of his values as percent of cellular carbon or as ATP per cell.

For algal samples the ratio of ATP to chlorophyll seems to have some meaning. As shown in Table 4, the ATP to chlorophyll ratio is relatively constant even under different nutritional and physical conditions. Using the ratios of ATP to chlorophyll seems more reliable than using the ATP to dry weight ratio, especially for natural populations. Even in unialgal cultures the ATP to dry weight ratio may not be reliable. Lee *et al* (1971), working with batch cultures of *Selenastrum capricornutum*, only calculated the ATP to dry weight ratio up to four days of batch growth. After four days they observed cell debris and dead cells, making the dry weight meaningless.

Lehninger (1971) estimated that the turnover time of the *E. coli* ATP pool is only a fraction of a second. Forrest and Walker (1965) have found the turnover time in *Streptococcus faecalis* to be about 5 seconds. In light of the rapid turnover rate of cellular ATP, many researchers have studied the response of cellular ATP to different environmental conditions.

Forrest (1965) showed that the ATP pool in *Strep faecalis* increased after addition of glucose and later returned to its endogenous level when the substrate was



used up. Working with *E. coli*, Wimpenny (1967) also noted an increase in ATP pool size when glucose or pyruvate was added.

Patterson *et al* (1969) and Patterson (1970), working with bench scale activated sludge units, studied the response of the ATP pool to changes in the incubation temperature and pH, to extended anaerobiosis, to starvation and enrichment, and to inhibition by nickel, chromium, and chromate. He also studied the ATP response of activated sludge to additions of toxic materials including mercury, copper, and cyanide. Maximum ATP pool occurred in the pH range of 7.5 to 8.0, the normal operating range of the units used in his research. Within 15 minutes after addition of a lethal dose of mercuric chloride to activated sludge, no ATP was detected.

Brezonik and Patterson (1971), reporting on the effects of environmental stress on ATP in activated sludge, observed increased ATP pool levels following addition of substrate. They noted, however, that previous studies on endogenous ATP pool indicated that only a small fraction of activated sludge is viable. Patterson *et al* (1970) reported 15 to 20 percent viability in a full scale activated sludge unit and 35 to 40 percent in a laboratory unit. Brezonik and Patterson questioned whether the increased ATP pool levels following substrate addition reflect an increase in cell population or an increase in ATP/cell. Recalculating the ATP response on a viable fraction basis, Brezonik and

Patterson, showed that substrate addition can effect an increase in ATP per cell.

Coomb *et al* (1967a) subjected cultures of the marine diatom *Cylindrotheca fusiformis* to 24 hours of darkness followed by re-illumination at a high light intensity and addition of silicon. During the time of silicon uptake, the ATP pool size decreased with a subsequent increase when the silicon uptake ceased. Although this indicated that ATP participates in silicon metabolism of cell wall formation, the presence of other cellular activities such as cell division could also account for the decrease in ATP pool size. Coomb *et al* (1967b) also studied the same phenomena using *Navicula pelliculosa*, a diatom which does not divide when subjected to a period of silicon starvation. Addition of silicon induces synchronous uptake of silicon and wall formation. When silicon was added, a rapid temporary increase in the ATP pool occurred followed by a sharp decrease. The ATP pool increased slowly throughout the remainder of the silicon uptake period. Similar changes occurred in a synchronous culture kept in the dark.

Santarius and Heber (1965) studied the changes in ATP, ADP, AMP and inorganic phosphate in leaf cells. Leaves were exposed to light and dark, killed and fractionated into a chloroplastic and a residual fraction. When the chloroplast was exposed to light, the ATP increased rapidly and it decreased rapidly when the chloroplast was placed in the dark. The ATP pool responded

in a similar manner in the cytoplasmic fraction. The ADP change was opposite to the ATP change and the AMP change followed that of ADP. From these results, Santarius and Heber concluded that the controlling factor for inhibition of glycolysis and respiration by light is the increased ratio of ATP to ADP rather than a drop in the orthophosphate concentration.

Because of the energy balance occurring within cells, ATP seems to be relatively constant under normal environmental conditions. This constant ATP pool may allow the use of ATP as a measure of microbial biomass. On the other hand, the rapid turnover time and sensitivity of ATP to environmental stress may permit the use of ATP analysis as a rapid bioassay method.

## CHAPTER IV. EXPERIMENTAL METHODS AND MATERIALS

### Analytical Techniques

The procedures used in this study were basically those commonly used to analyze the biological and chemical constituents of aquatic systems. Analysis of ATP is discussed separately in Chapter V.

Chemical oxygen demand (COD) was measured using the dichromate reflux method described in Standard Methods (APHA 1971). Distilled water blanks were analyzed simultaneously. Samples were not filtered prior to analysis thus the reported values represent combined soluble and suspended particulate chemical oxygen demand.

The dry weight of the algal cultures was determined using the gravimetric procedure recommended in the Provisional Algal Assay Procedure (Joint Industry-Government Task Force on Eutrophication 1969). A measured portion of algal suspension was filtered through a tared type AA millipore filter with a 0.80 micron pore size. The filters were dried for several hours at 90° C in an oven, then they were placed in a desiccator to cool. The filters were weighed on a Mettler balance.

Turbidity was measured in Jackson Turbidity Units (JTU) using a Hach Turbidimeter. Absorbance was measured as described in the Provisional Algal Assay Procedure (Joint Industry-Government Task Force on Eutrophication 1969). A Bausch and Lomb Spectronic 20 spectrophotometer at a wavelength of 600 nm was used. Conductivity was measured as described in Standard Methods (APHA 1971) using a Beckman conductivity bridge.

Ortho-phosphate was measured by the single reagent molybdenum blue method of Murphy and Riley (1962) adapted to the Technicon Autoanalyzer. Total phosphorus was measured by autoclaving samples at 15 psi for one hour in the presence of potassium per sulfate and sulfuric acid. The phosphorus concentration was measured using a Klett-Sumerson photoelectric colorimeter.

The modified brucine method of Jenkins and Medsker (1964) adapted to the Technicon Autoanalyzer was used to measure nitrate. This procedure was essentially identical to the automated method described by Kahn and Brezenski (1967).

Alkalinity was measured by potentiometric titration to pH 4.5 with 0.01 N H<sub>2</sub>SO<sub>4</sub> as described in Standard Methods (APHA 1971).

Chlorophyll *a* was extracted and measured by the procedure originally described by Richards and Thompson



(1952) and Crietz and Richards (1955). The equations of Parsons and Strickland (1963) were employed to calculate the chlorophyll *a* concentration.

A modification of the carbon-14 procedure developed by Steeman-Nielsen (1951) was used to measure primary production. Rather than inoculating 300 ml samples with 5 micro-curies of  $^{14}\text{C}$ , 30 ml samples were inoculated with 0.5 micro-curie of  $^{14}\text{C}$ . After incubation for three hours in a constant temperature light box, aliquots were filtered, dried, and counted in an internal proportional counter. A method study was performed to determine the statistical accuracy of this modified procedure. Replication of this procedure yields a relative standard deviation of five percent.

Alkaline phosphatase activity of the algal cultures was measured using a modification of the method of Fitzgerald and Nelson (1966). Instead of using Gorham's medium as suggested, PAAP medium was used since all algal cultures were grown on PAAP medium.

Cell numbers were determined using a hemocytometer.

#### Experimental Materials

Four laboratory batch unialgal cultures were operated for a period of seven months. The following

algal species were grown: *Anabaena flos-aquae*,<sup>1</sup> *Microcystis aeruginosa*,<sup>1</sup> *Selenastrum capricornutum*,<sup>1</sup> and *Chlorella* sp. Initially, chemostat units were set up using *Anabaena flos-aquae*, *Selenastrum capricornutum*, and *Navicula minima*,<sup>2</sup> a diatom. A detention time of three days was maintained by feeding three liters of PAAP medium to the nine liter chemostats. A gradual wash-out of the algal cells occurred, necessitating a change in detention time from three days to nine days. After one month of operation all three chemostats were contaminated with *Chlorella* and new culture material had to be ordered from the National Eutrophication Research Center in Oregon and from the Starr Collection in Indiana. At this time a decision was made to change to batch cultures.

Batch unialgal cultures of *Anabaena*, *Microcystis*, *Selenastrum*, and *Navicula* were grown. However, *Navicula* continued to be contaminated with *Chlorella*. Finally, *Navicula* cells declined in numbers and had to be discarded. A unialgal culture of *Chlorella* was substituted for the diatom.

The batch units contained four liters of culture material which were kept well-mixed with magnetic stirrers. Gro-Lux<sup>3</sup> fluorescent lamps were used to illuminate the

---

<sup>1</sup>Obtained from the National Eutrophication Research Program, Pacific Northwest Water Laboratory, Corvallis, Oregon 97350.

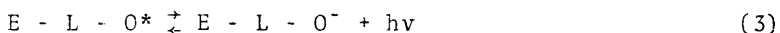
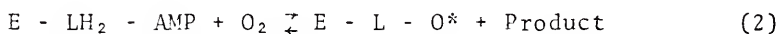
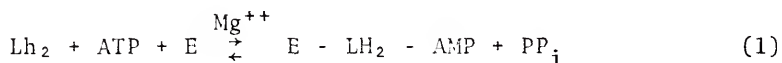
<sup>2</sup>Obtained from Dr. R. Starr, Dept. of Botany, Univ. of Indiana, Bloomington, Indiana.

<sup>3</sup>Sylvania Electric Products Inc., Danvers, Mass.

cultures. *Selenastrum* and *Chlorella* received about 350 ft-c while *Anabaena* and *Microcystis* received about 200 ft-c. Each day 200 ml of culture material was removed from each unit and 200 ml of fresh PAAP medium was added. The PAAP medium was sterilized before it was fed to the units. Because of the poor buffer capacity of PAAP medium, the pH of the cultures often increased to a pH of 10 or greater, indicating a carbon-limiting condition. Carbon dioxide was occasionally bubbled into the units to lower the pH and add carbon. Frequent visible changes in the algal pigments of the units occurred, especially when the pH increased. At no time was a steady-state condition obtained.

## CHAPTER V. METHODS OF ATP ANALYSIS

McElroy (1947) found that luminescence in fireflies has an absolute requirement for ATP. Light production by firefly lantern extract (FLE) depends on the presence of luciferin, the enzyme luciferase, oxygen, magnesium ions, and ATP (McElroy 1947; McElroy and Strehler 1954). Hastings (1968) has reviewed the biochemistry of luminescent reactions in detail. Equations 1-3 illustrate the chemical reactions involved in the firefly light production process.



Luciferin and ATP, catalyzed by luciferase, react to form an enzyme-luciferin-adenosine monophosphate complex (luciferyl-adenylate) and inorganic pyrophosphate (PP<sub>i</sub>). Rapid oxidation to oxyluciferyl-adenylate occurs (Equation 2). This is an excited state and is immediately followed (Equation 3) by the release of a quantum of light (Hopkins, Selinger *et al* 1968). One light quantum is emitted for each luciferin molecule oxidized (Selinger and McElroy 1959, 1960).

The mixing of ATP with firefly lantern extract results in a flash of light which rapidly declines to a uniform level. Addition of arsenate or phosphate buffer to the firefly reaction decreases the initial light flash and produces an intermediate level of light which decays exponentially with time. Arsenate buffer is now routinely added to many commercially available firefly lantern extracts used for ATP analysis.

ATP is the only nucleoside triphosphate that will react with purified extracts of firefly lantern to produce luminescence. The crude extracts of firefly lantern, however, contain transphosphorylase enzymes, resulting in light emission in the presence of high energy phosphate molecules other than ATP (Balfour and Samson 1959). To determine which high energy phosphate molecules might cause light emission with firefly extract, Holm-Hansen and Booth (1966) tested thirteen intracellular compounds including adenylic acid, guanylic acid, cozymase, glucose-1-phosphate, fructose-1-6-diphosphate, phosphocreatine, adenosine, thiamine pyrophosphate, coenzyme-A, sodium glass, adenosine diphosphate (ADP), cytidine-5-triphosphate (CTP) and inosine-5-triphosphate (ITP). Of these compounds, only ADP, CTP, and ITP affected light emissions. The amount of light resulting from addition of ADP was less than 1 percent of an equivalent amount of ATP. Both CTP and ITP stimulated light production equivalent to that of ATP. Franzen and

Binkley (1961) reported cellular ATP values 10 to 25 per cent higher using the firefly method than when determined by chromatographic techniques. The fact that some light emission may be due to nucleoside triphosphates other than ATP is probably not significant because of the relative abundance of cellular ATP with respect to these other compounds.

Since one quantum of light is emitted for each molecule of ATP that is hydrolyzed (Seliger and McElroy 1960), the amount of ATP in a sample is directly proportional to the total amount of light emitted by the enzyme mixture. ATP was measured based on this principle.

Lyophilized aqueous extracts of firefly lantern<sup>1</sup> were stored desiccated at  $-20^{\circ}$  C until used. Each vial contains the extract from 50 mg of firefly lanterns with magnesium arsenate buffer added. The contents of each vial was rehydrated with 35 ml of deionized distilled water. After standing at room temperature for 1/2 hour, the enzyme mixture was filtered (Whatman No. 3) and the filtrate incubated in ice water for 3 to 4 hours. The crystalline disodium salt of ATP was used to prepare standard solutions. A stock solution was prepared in tris buffer (0.025 M, pH 7.75), and poured into individual test tubes which were capped and stored at  $-20^{\circ}$  C until needed. Storage for as long as two months showed no change in standard ATP

---

<sup>1</sup>FLE-50, Sigma Chemical Company, St. Louis, Missouri.

concentration (Patterson 1970). When ATP standards were needed a test tube of stock solution was thawed and diluted with tris buffer to the desired concentrations.

Samples to be analyzed for ATP were filtered through 47-mm membrane filters (pore size 0.80  $\mu$ ). As soon as no liquid remained above the filter, the filter was quickly removed and placed in a test tube containing 9 ml of boiling tris buffer (0.025 M, pH 7.75). The test tube was held in a boiling water bath for 10 minutes, with occasional shaking to disperse clumped cells. The test tube was rapidly cooled and brought up to a volume of 10 ml with additional tris buffer. The test tube was centrifuged (10 minutes at 2,250 rpm) to bring down cell debris, and the supernatant was poured into a clean test tube. The test tube was then frozen and stored at  $-20^{\circ}$  C for future analysis.

This method of ATP extraction depends upon bringing all of the ATP into solution. Other extraction methods have been used. Holm-Hansen and Booth (1966) tested ATP extraction using perchloric acid, boiling ethanol, boiling water, and boiling tris buffer. Perchloric acid extraction inhibited the luciferase reaction because of its neutral pH. Extraction by the last three solvents all gave satisfactory results. Forest (1965) used sulfuric acid to extract ATP and D'Eustachio and Levin (1967) used sonic disruption to extract ATP. Patterson (1970) compared sonication to boiling tris extraction and found it was

equally effective in extracting ATP. DuPont de Nemours Company (Inc.) recommends ATP be extracted with either butanol or dimethylsulfoxide (DMSO) for use with their Luminescence Biometer.<sup>2</sup>

A sensitive method of measuring light emission is required in the ATP assay. Holm-Hansen and Booth (1966) used a photomultiplier tube and amplifier connected to a recorder. Others (Forrest 1965; Lyman and DeVincenzo 1967) have also used similar instrumentation. Another instrument used to measure ATP is the DuPont Luminescence Biometer. The DuPont Luminescence Biometer consists of a photomultiplier tube and solid-state electronics that converts the analog signal to a digital readout proportional to the maximum intensity of the light flash. The instrument is calibrated for each reaction mixture so that the ATP concentration is read out directly. Purified luciferin-luciferase is used without arsenate buffer to give a high initial light flash and a rapid luminescent die-off. At the start of this research the Biometer was used to measure ATP concentration. During a two month investigation period difficulties were encountered with calibration of the instrument and read-out reproducibility. The Biometer is also sensitive to fluctuations in line voltage entering the instrument. Lack of reproducibility was probably caused by a combination of the complex solid-state electronics

---

<sup>2</sup>DuPont 760 Luminescence Biometer, DuPont de Nemours and Company, Wilmington, Delaware.



and the sample size and injection procedure. The DuPont procedure calls for injection of a 0.01 ml aliquot of sample into a vial containing 0.1 ml of enzyme mixture. Although a precision microliter syringe was used to inject the sample, it is doubtful whether exactly 0.01 ml was delivered each time at the same rate. Small droplets would sometimes cling to the needle point, indicating a lack of precision in the injection procedure. In general, the Biometer was found to be an unsatisfactory method of measuring ATP.

Conventional liquid scintillation spectrometers have been used to measure luminescence (Cole, Wimpenny, *et al.* 1967; Patterson 1970). A liquid scintillation spectrometer<sup>3</sup> was used in this study. Instrument gain was set at 53, with a window opening of 50 to 1,000. The spectrometer was set in a repeat count mode, with each 6-second counting interval separated by a 7.5 second data print-out sequence.

Prior to each ATP analysis the background light emission from the luciferin-luciferase firefly extract was measured. Exactly 1.0 ml of the enzyme preparation was pipetted into a scintillation vial. Normal background emission was 10 to 20 counts in the 6-second counting interval. Patterson (1970) found that higher counts indicated glassware contamination. All glassware was washed

---

<sup>3</sup>Packard Tri-Carb Model 2002, Packard Instrument Company, Downers Grove, Illinois.

in an automatic dishwasher using hot soapy water, a hot rinse, and a final rinse with distilled water. Following this, the glassware was boiled for at least one hour in an acid bath and rinsed three times in deionized, distilled water.

After determining the background emission, exactly 0.5 ml of ATP standard or sample was pipetted into the enzyme mixture and the vial was swirled to thoroughly mix the contents. Exactly 11.0 seconds after ATP addition, the vial was placed into the scintillation counting chamber. This procedure was followed because of the exponential decay of the luminescence, necessitating careful control over the addition of ATP to the enzyme mixture and initiation of the counting sequence.

Patterson (1970), using the scintillation counter to measure ATP, analyzed the data graphically because of the random variability of any 6-second count. The 6-second emission counts were plotted on semi-log paper versus elapsed reaction time. Good correlation between counts and standard ATP concentrations were observed for reaction times of one minute or more. The line of best fit to the data points was extrapolated back to the initial data point and the count at one-minute was read from this line. This graphical technique was used in this research. Figure 4 shows luminescence decay curves for several ATP concentrations.

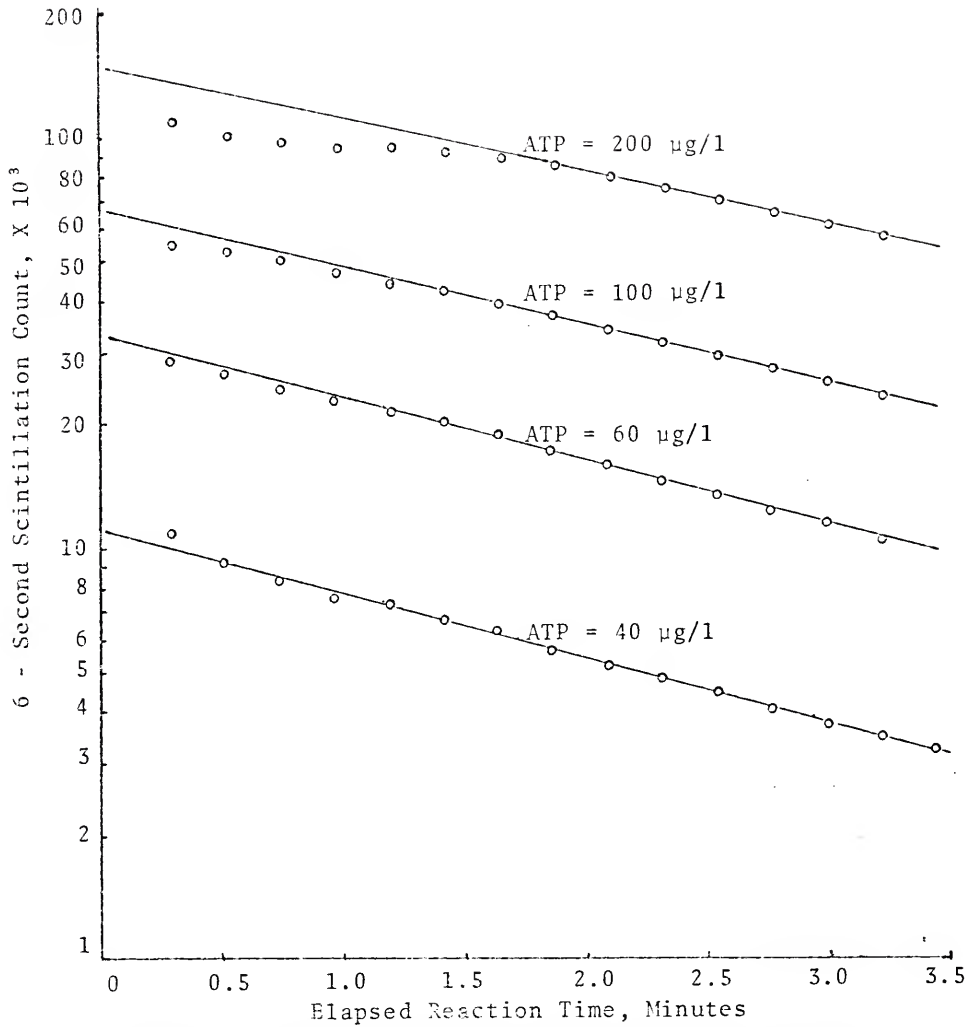


FIGURE 4. LUMINESCENCE DECAY OR ATP FIREFLY LANTERN EXTRACT REACTION WITH TIME (Patterson 1970)

The effects of aging on both ATP standard solutions and firefly lantern extract were studied (Patterson 1970). As shown in Table 5, the background emission of the luciferin-luciferase mixture decreased rapidly during a 24-hour incubation period. Two sets of ATP standards were also made up: one set was frozen, then thawed just prior to analysis; the other was held in an ice water bath with the firefly mixture. The light emission from the freshly thawed standards decreased with the time of incubation, indicating a slight loss of enzyme activity over the 24-hour period. Light emission from the ATP solution kept in an ice bath decreased rapidly, indicating a loss of ATP from the standard solutions during the incubation period. In light of these results, standards were always measured immediately after preparation, and samples were thawed just prior to analysis.

The effect of tris buffer temperature on ATP extraction was also studied by Patterson (1970). He found (Figure 5) that a temperature reduction of a few degrees below boiling caused a significant decrease in the amount of ATP extracted. For this reason, all extractions were performed in a rapidly boiling water bath. However, it was often found that water which appeared to be boiling actually had a temperature of 90-95°C. Thus a thermometer was always kept in the boiling tris buffer to assure that the temperature remained at 100°C.

TABLE 5

EFFECT OF INCUBATION ON FIREFLY EXTRACT  
AND ATP LIGHT EMISSION (6-SEC COUNT)  
(Patterson 1970)

Extract Age, Hours	Extract Background Counts	Six Second Count			
		Fresh ATP, $\mu\text{g/l}$		Aged ATP, * $\mu\text{g/l}$	
		20	100	20	100
2	355	11,200	118,000	500	500
6	153	7,500	121,000	430,000	418,000
12	63	-----	-----	-----	383,000
24	15	3,350	85,000	400,000	56,200
				25	324,000

\* ATP age was one hour less than extract age

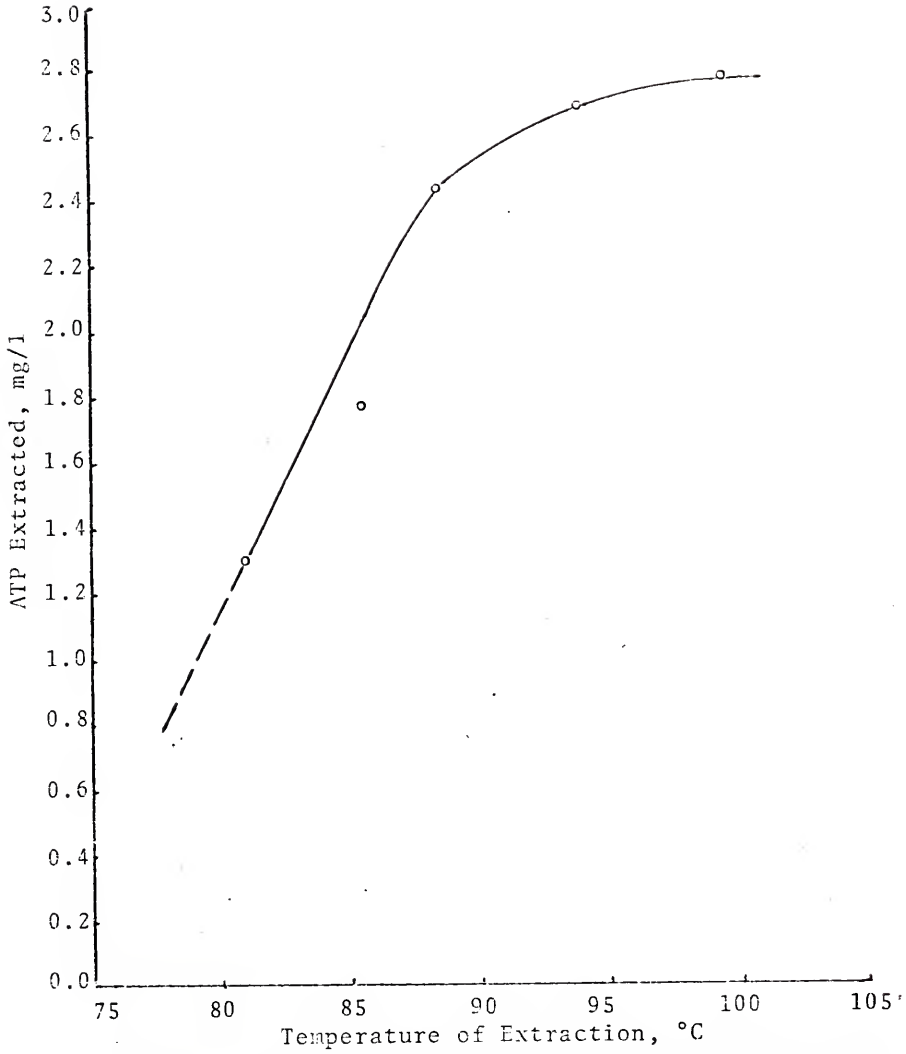


FIGURE 5. DEPENDENCE OF QUANTITATIVE ACTIVATED SLUDGE  
ATP EXTRACTION ON TEMPERATURE OF EXTRACTION SOLUTION  
(Patterson 1970)

Rhodes and McElroy (1958) investigated the sensitivity of the firefly reaction to pH and found that luminescence increased rapidly from pH 2.0 to 3.8, decreased from pH 2.8 to 5.8, and increased from pH 4.8 to 7.0. The rate of light emission is much more rapid at pH 7.6 than at pH 9.4 (Seliger and McElroy 1960). Holm-Hansen (1966) used a reaction mixture buffered at pH 7.75. For this research, tris buffer at a pH of 7.75 was used.

Figure 6 shows a typical standard ATP curve. Depending on the enzyme age and concentration, the standard curve may be linear or non-linear. Results of recent measurements of ATP standards indicate that plotting the points on log-log graph paper gives a straight line of best fit.

A sample of *Selenastrum* was diluted into five portions and the ATP of each portion was measured. This experiment was run to determine whether the sampling procedure was valid. As shown in Figure 7, excellent correlation between ATP content and *Selenastrum* concentration was observed.

Using the method of ATP analysis described in this chapter, replication of ATP samples gave results with a relative standard deviation of less than 5 percent.

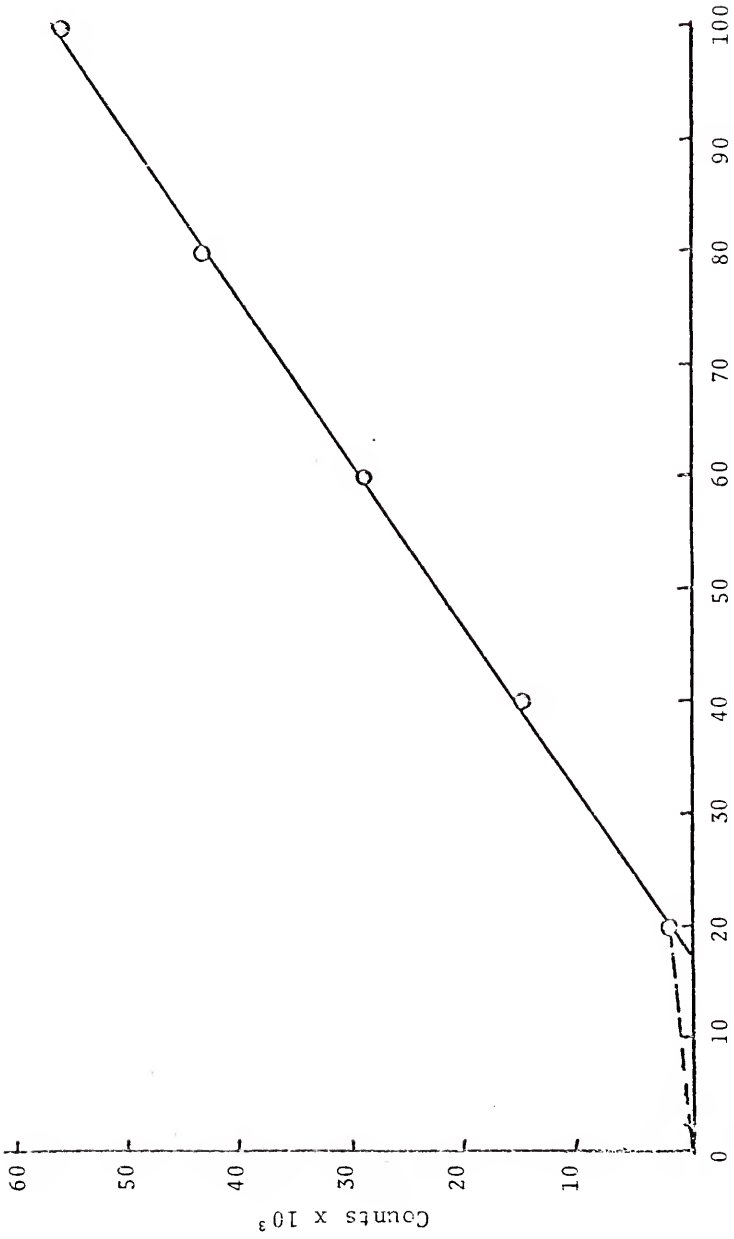


FIGURE 6. STANDARD ATP CURVE



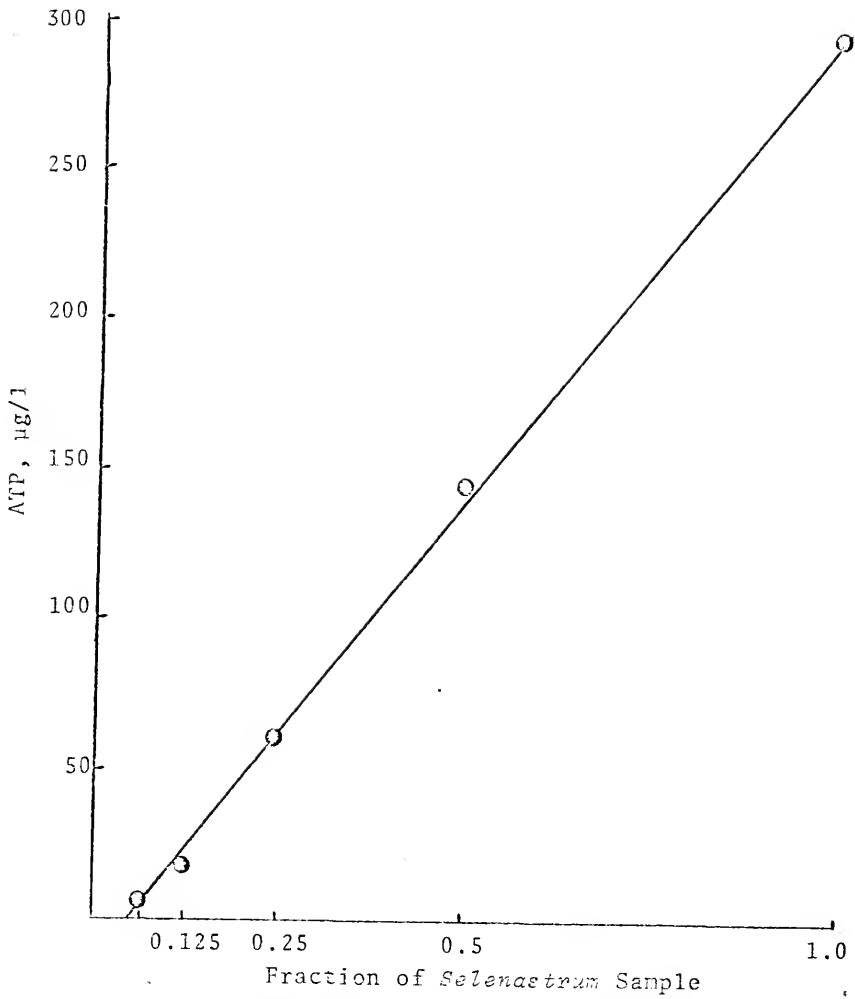


FIGURE 7. CORRELATION OF ATP WITH *SELENASTRUM* CONCENTRATION

## CHAPTER VI. EXPERIMENTAL RESULTS

The study of ATP in phytoplankton was divided into two phases. The first phase consisted of monitoring the ATP pool in batch grown unialgal cultures and natural lake populations. In the second phase, samples from the unialgal cultures and from lake waters were subjected to various environmental conditions such as pH, light and addition of nutrients and toxic substances. The ATP response to these conditions was measured.

### ATP - Biomass Results

Batch cultures of *Anabaena flos-aque*, *Selenastrum capricornutum*, *Microcystis aeruginosa*, and *Chlorella* sp. were grown in PAAP media for a period of approximately three months. Twice each week a 200 ml sample was taken from each culture and analyzed for ATP and other biomass parameters including chlorophyll *a*, dry weight, and absorbance. These experiments were performed to determine whether the cellular ATP concentrations in batch unialgal cultures would correlate with current biomass parameters (chlorophyll *a*, dry weight, etc.). Samples of phytoplankton from local lakes were also collected and analyzed for ATP, chlorophyll *a*, dry weight, and absorbance. The correlation of ATP with current biomass

parameters was determined to test the usefulness of ATP measurements in natural phytoplankton populations. No attempt was made to quantitatively measure the bacteria population in the batch cultures or the lake samples.

#### ATP vs. Chlorophyll *a*

Figures 1 to 3 present ATP vs. chlorophyll *a* values for the batch algal cultures. The slopes of the linear regression lines for *Selenastrum* and *Chlorella* are similar; thus Figure 8 contains the data for both these organisms. Although the slopes of the linear regression lines are similar for *Anabaena* and *Microcystis*, the ATP and chlorophyll *a* values are plotted separately for reasons of clarity and because of the much larger range of chlorophyll *a* values observed in the *Anabaena* culture.

Although there is a wide degree of scatter about the regression lines for all algal species, a definite relationship exists. Correlation coefficients for ATP vs. chlorophyll *a* concentrations are relatively high as shown in Table 6. Correlation is a measure of the degree to which variables vary together, that is, it is a measure of the intensity of association. The sample correlation coefficient,  $r$ , given in Table 6 is an estimate of the population coefficient,  $\rho$ . Table 6 gives the 95 percent and 99 percent confidence intervals for the population correlation coefficient. Note that a correlation does not exist if the population correlation coefficient is zero. Thus, to be significant

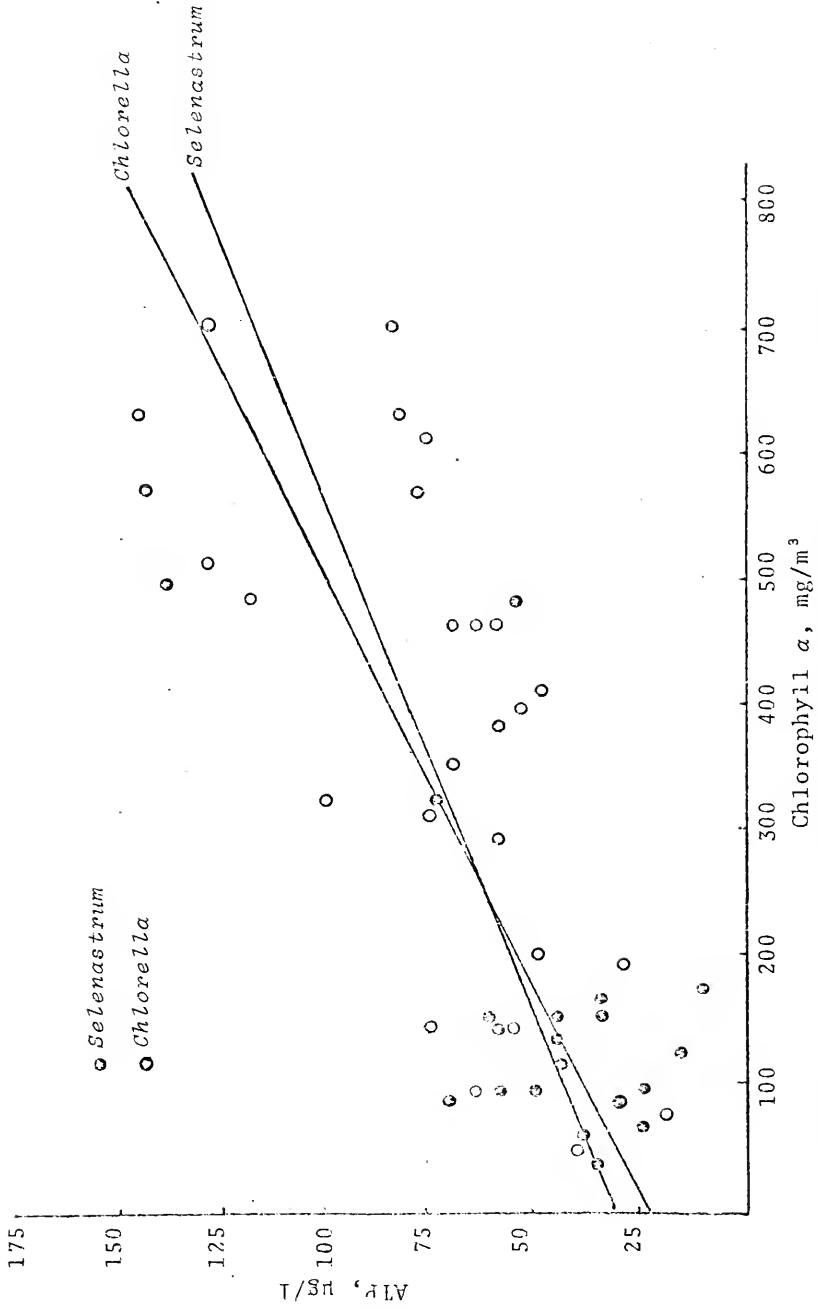


FIGURE 8. ATP VS. CHLOROPHYLL A OF SELENASTRUM AND CHLORELLA

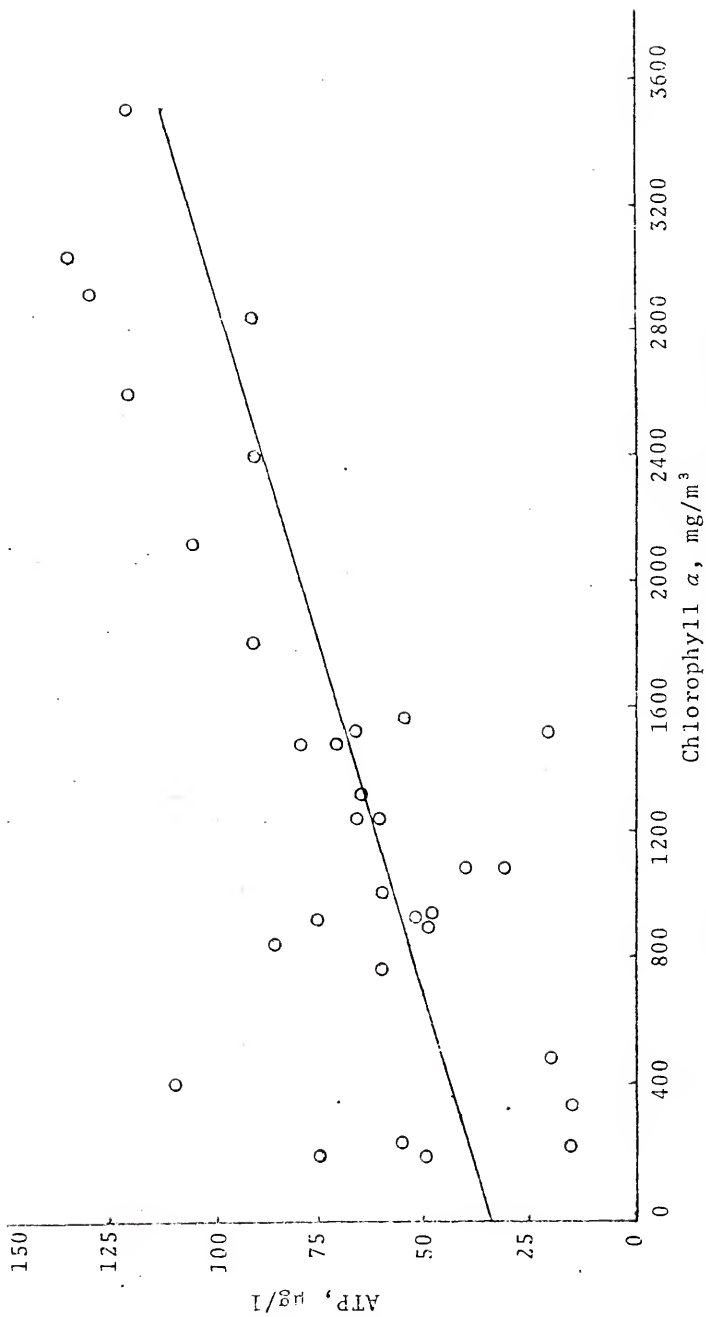


FIGURE 9. ATP VS. CHLOROPHYLL A FOR ANABAENA

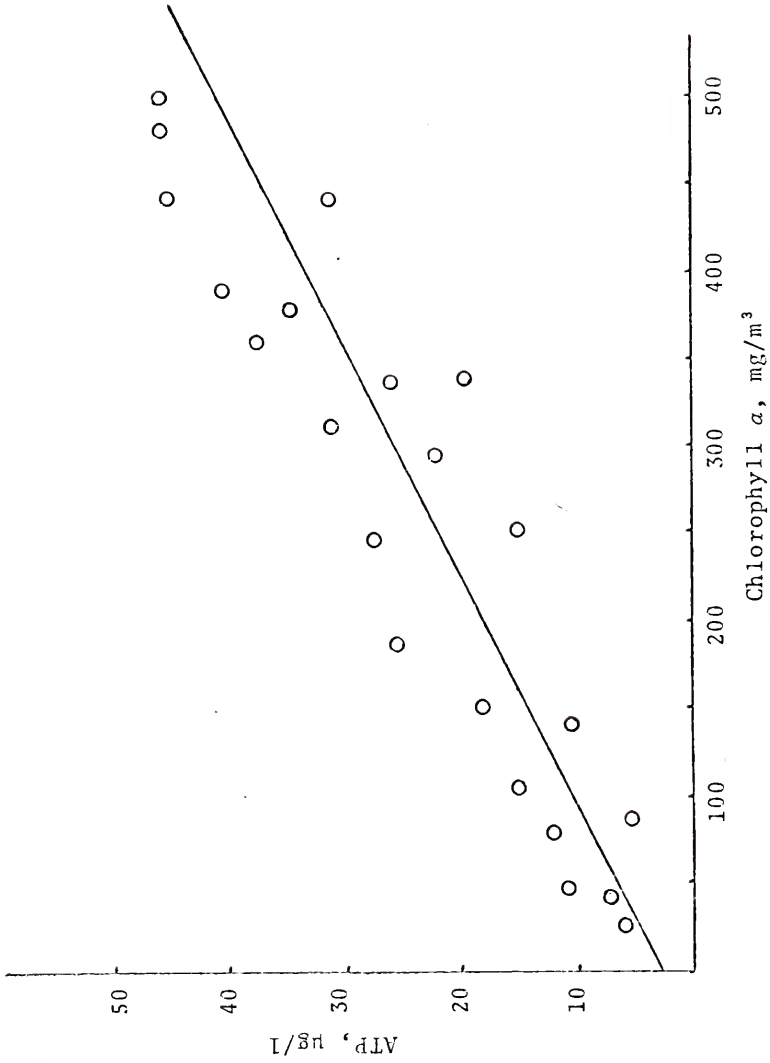


FIGURE 10. ATP VS. CHLOROPHYLL *A* FOR *MICROCYSTIS*

TABLE 6

ATP VS. CHLOROPHYLL A CORRELATION COEFFICIENTS FOR  
UNIALGAL CULTURES

Algae	Sample Correlation Coefficient, r	95% Confidence Interval, p	99% Confidence Interval, p
<i>Anabaena</i>	0.707	0.45 - 0.83	0.35 - 0.87
<i>Selenastrum</i>	0.539	0.15 - 0.76	0.01 - 0.82
<i>Microcystis</i>	0.683	0.36 - 0.84	0.28 - 0.87
<i>Chlorella</i>	0.725	0.45 - 0.87	0.31 - 0.89

at a particular confidence level, the confidence interval must not contain zero. As shown in Table 6, the sample correlation coefficients for all algal species are significant at the 99 percent confidence level, but the correlation for *Selenastrum*, which had the lowest correlation ( $r = 0.539$ ) is barely significant at the 99 percent confidence level since its confidence interval approaches zero at the lower limit.

To study further the relationship between ATP and chlorophyll *a*, a dense culture of *Microcystis* was diluted into 10 portions, incubated for 24-hours, and analyzed for ATP and chlorophyll *a*. A nearly perfect linear relationship was obtained (Figure 11). These results demonstrate a strict ATP to chlorophyll *a* relationship in the absence of varying conditions of pH, light intensity, and nutrient state.

To determine the relationship of ATP to chlorophyll *a* in natural lake phytoplankton, measurements were made on three Florida lakes: Bivens Arm, Newnan's Lake, and Anderson-Cue Lake. These lakes were selected to provide a wide range of biomass values. Bivens Arm is a hypereutrophic lake, Newnan's Lake is eutrophic, and Anderson-Cue Lake is oligotrophic. Chemical and biological characteristics of these lakes are given in Table 7. Results of the lake measurements are shown in Figure 12, and the correlation coefficients and confidence intervals are given in Table 8. Correlation of ATP with chlorophyll *a* for all



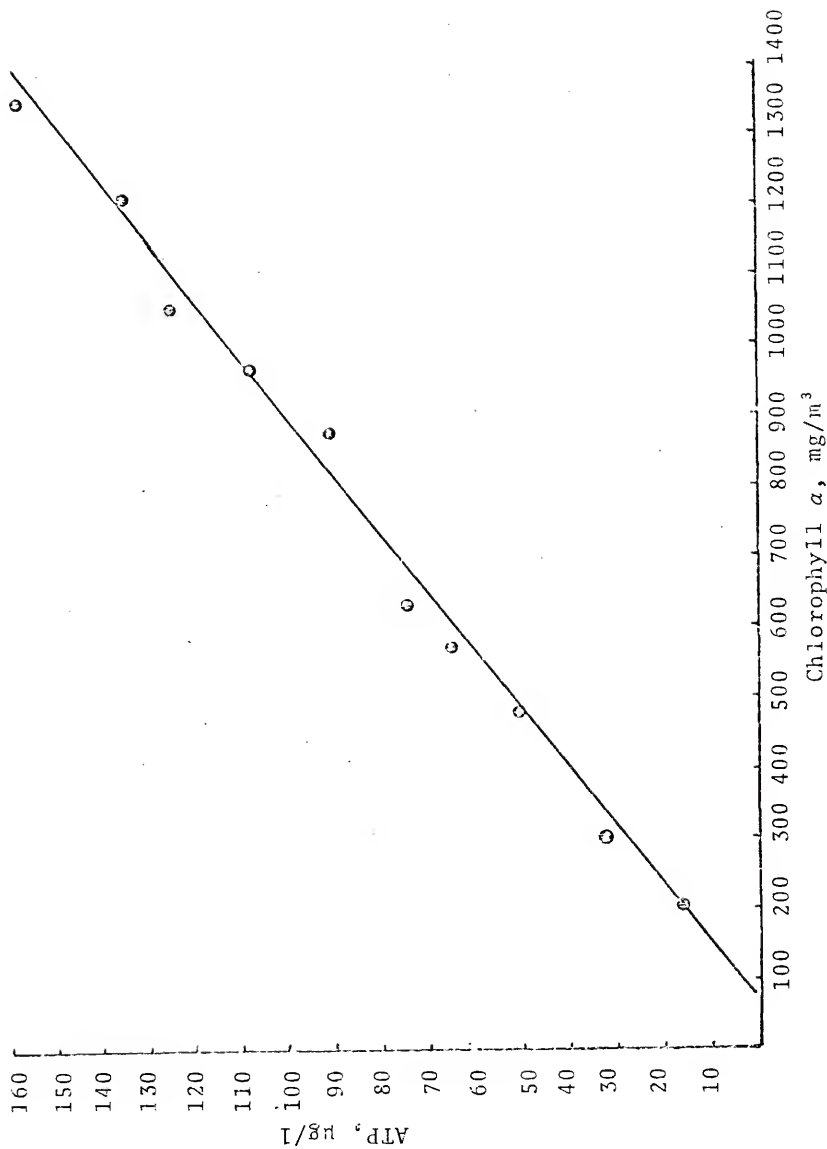


FIGURE 11. ATP VS. CHLOROPHYLL A OF DILUTED MICROCYSTIS

TABLE 7  
 CHEMICAL AND BIOLOGICAL CHARACTERISTICS  
 OF EXPERIMENTAL LAKES<sup>1</sup>

	Bivons Arm	Newnan's Lake	Andersen-Cue
pH	8.99	7.78	4.76
Acidity (mg/l CaCO <sub>3</sub> )	0.0	1.01	2.68
Alkalinity (mg/l CaCO <sub>3</sub> )	105.0	7.3	0.0
Specific Conductance (umho cm <sup>-1</sup> )	253.8	59.8	38.2
Color (mg Pt/l at pH 8.3)	42.1	188.9	3.0
Total Organic Nitrogen (mg/l)	1.87	1.41	0.28
NH <sub>3</sub> (mg/l)	0.19	0.36	0.14
NO <sub>3</sub> (mg/l)	0.05	0.07	0.09
Ortho-PO <sub>4</sub> (mg/l)	0.225	0.012	0.002
Total-PO <sub>4</sub> (mg/l)	0.546	0.110	0.018
Turbidity (JTU)	10.2	4.2	1.0
Primary Production (mg C/m <sup>3</sup> -hr)	251.7	102.5	1.1
Chlorophyll <i>a</i> (mg/m <sup>3</sup> )	56.03	47.39	3.17

<sup>1</sup> Average of data collected from June 1968 to April 1970 (Shannon 1970)

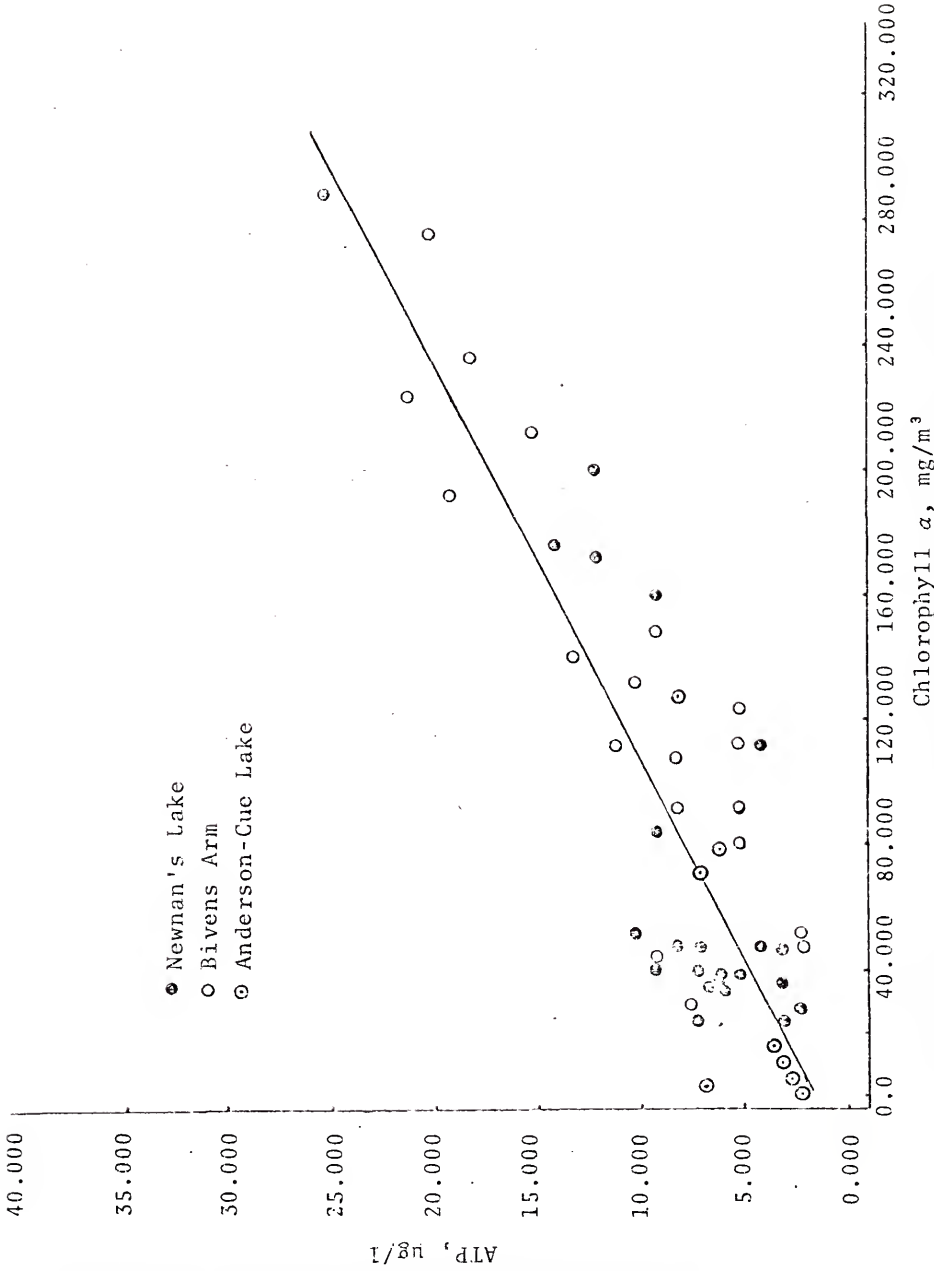


FIGURE 12. ATP VS. CHLOROPHYLL  $a$  FOR LAKE PHYTOPLANKTON

TABLE 8

ATP VS. CHLOROPHYLL A CORRELATION COEFFICIENTS FOR  
LAKE PHYTOPLANKTON

Lake	Sample Correlation Coefficient, r	95% Confidence Interval, $\rho$	99% Confidence Interval, $\rho$
Bivens Arm	0.869	0.70 - 0.94	0.66 - 0.96
Newman's Lake	0.718	0.42 - 0.86	0.28 - 0.90
Anderson-Cue Lake	0.790	0.20 - 0.94	-0.10 - 0.97
All Lakes	0.830	0.73 - 0.92	0.68 - 0.94

three lakes is significant at the 95 percent confidence level. Bivens Arm and Newnan's Lake are significant at the 99 percent confidence level. However, the correlation is not very good for Anderson-Cue Lake alone as evidenced by its large confidence interval and its not being significant at the 99 percent confidence level. The poor correlation of ATP with chlorophyll *a* in Anderson-Cue is probably a result of the small range of values measured. The correlation coefficient of the combined lake values is 0.830. These high correlation values indicate the excellent association of ATP with chlorophyll *a* in lake phytoplankton.

The ATP to chlorophyll *a* ratios for the laboratory algal cultures and the lake samples were calculated and are given in Table 9. *Anabaena* and *Microcystis* have lower ATP to chlorophyll *a* ratios than *Selenastrum* and *Chlorella*. This could be explained by the fact that *Anabaena* and

TABLE 9  
ATP TO CHLOROPHYLL *A* RATIOS FOR ALGAL CULTURES  
AND LAKE PHYTOPLANKTON

	<u>ATP/Chlorophyll <i>a</i></u>
<i>Anabaena</i>	0.09
<i>Microcystis</i>	0.09
<i>Selenastrum</i>	0.35
<i>Chlorella</i>	0.29
Bivens Arm	0.09
Newnan's Lake	0.11
Anderson-Cue Lake	0.69
All Algae	0.20
All Lakes	0.25
Algae and Lakes	0.21

*Microcystis*, both blue-green algae, are procaryotic cells with a pigment system containing chlorophyll *a*, phycocyanin, and trace amounts of phycoerythrin. A very efficient energy transfer, approaching 100 percent, occurs from phycocyanin to chlorophyll *a* (Brock 1970). Thus the amount of chlorophyll *a* required by a blue-green alga may be less than that required by a green alga. In blue-green algae the photosynthetic pigments occur in organized internal membranes unlike eucaryotic algae where the pigments occur in membrane-bound chloroplasts. This difference in cellular organization could also account for the observed differences in the ATP to chlorophyll *a* ratio.

Bivens Arm and Newnan's Lake both have lower ATP to chlorophyll *a* ratios than Anderson-Cue Lake. One would almost expect the opposite, considering that Bivens Arm and Newnan's Lake probably contain a larger bacterial population than Anderson-Cue Lake. However, other factors must be considered. For instance, both Bivens Arm and Newnan's Lake contain large populations of blue-green algae while Anderson-Cue Lake does not. A large population of blue-green algae, according to the above data, would give a lower ATP to chlorophyll *a* ratio. Also, the much higher turbidity and color in Bivens Arm and Newnan's Lake probably induce a higher cellular chlorophyll content in order to utilize the subdued light entering the water.

Other experiments were performed to determine whether the ATP to chlorophyll *a* ratio varied with nutritional state and cell age. Aliquots from each of the unialgal cultures were transferred to flasks containing fresh media. One flask contained PAAP medium with all nutrients present, one contained PAAP medium without phosphorus, and one contained PAAP medium without nitrogen. The ATP to chlorophyll *a* ratio increased as the nutritional content of the medium decreased and the cell age increased (Figure 13). In most cases, the ATP to chlorophyll *a* ratio remained relatively constant for the first few days then increased. This resulted from an initial increase in chlorophyll *a* and ATP concentrations followed by a decrease in chlorophyll *a* as the cultures aged and the nutrient content of the medium decreased. Fogg (1965) observed that the cellular chlorophyll concentration decreased with cell age and nutrient deficiency. Thus, variations in the ATP to chlorophyll ratio may be a result of natural fluctuations in cellular chlorophyll rather than changes in cellular ATP.

#### ATP vs. Dry Weight

ATP and dry weight values of the batch algal cultures are shown in Figure 14. All four algal species have similar linear regression lines indicating a relatively constant association between ATP and dry weight. Correlation coefficients and their confidence intervals are given in Table 10. The sample correlation coefficients for *Anabaena*,

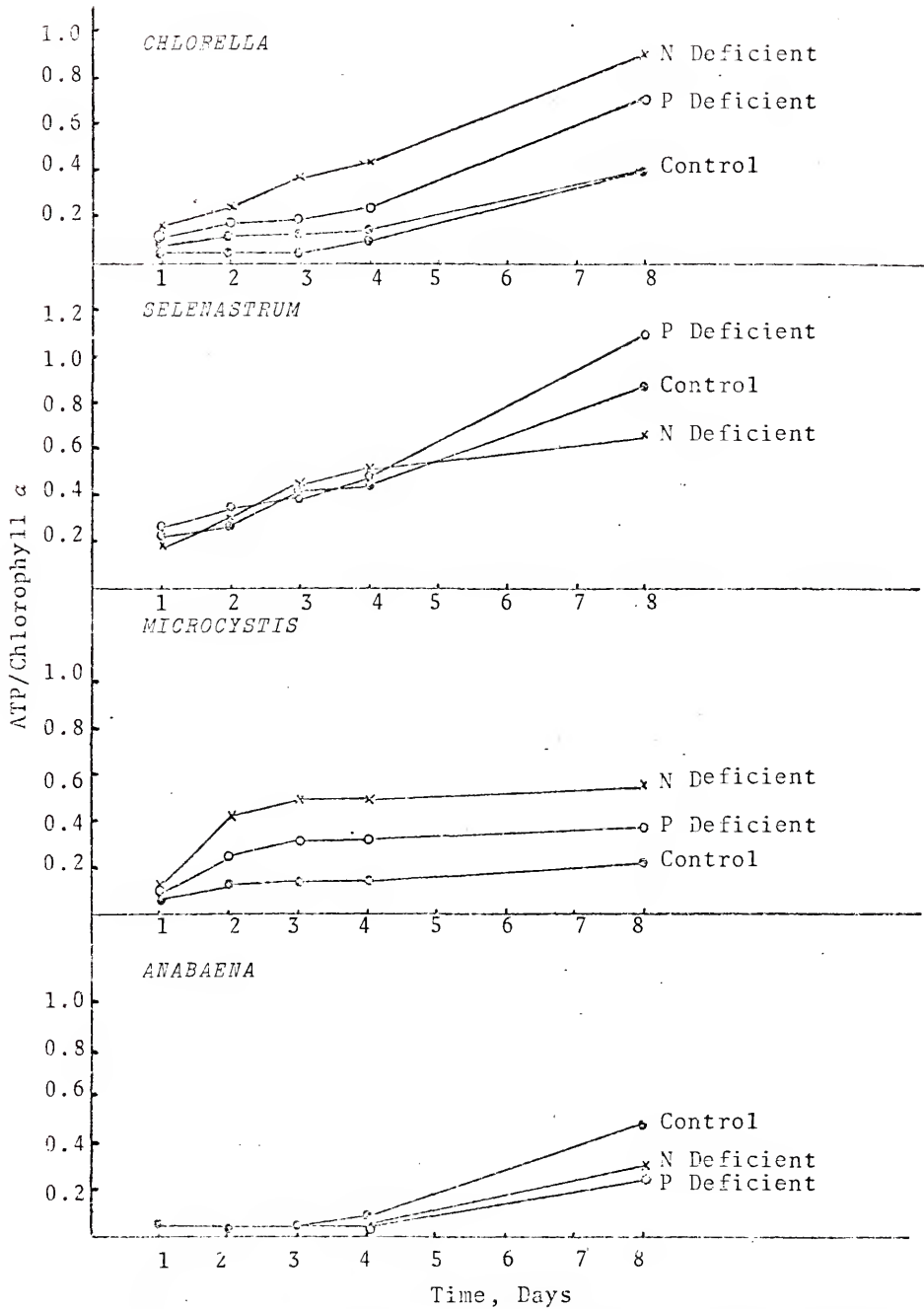


FIGURE 13. VARIATION OF ATP/CHLOROPHYLL A RATIO WITH CELL AGE



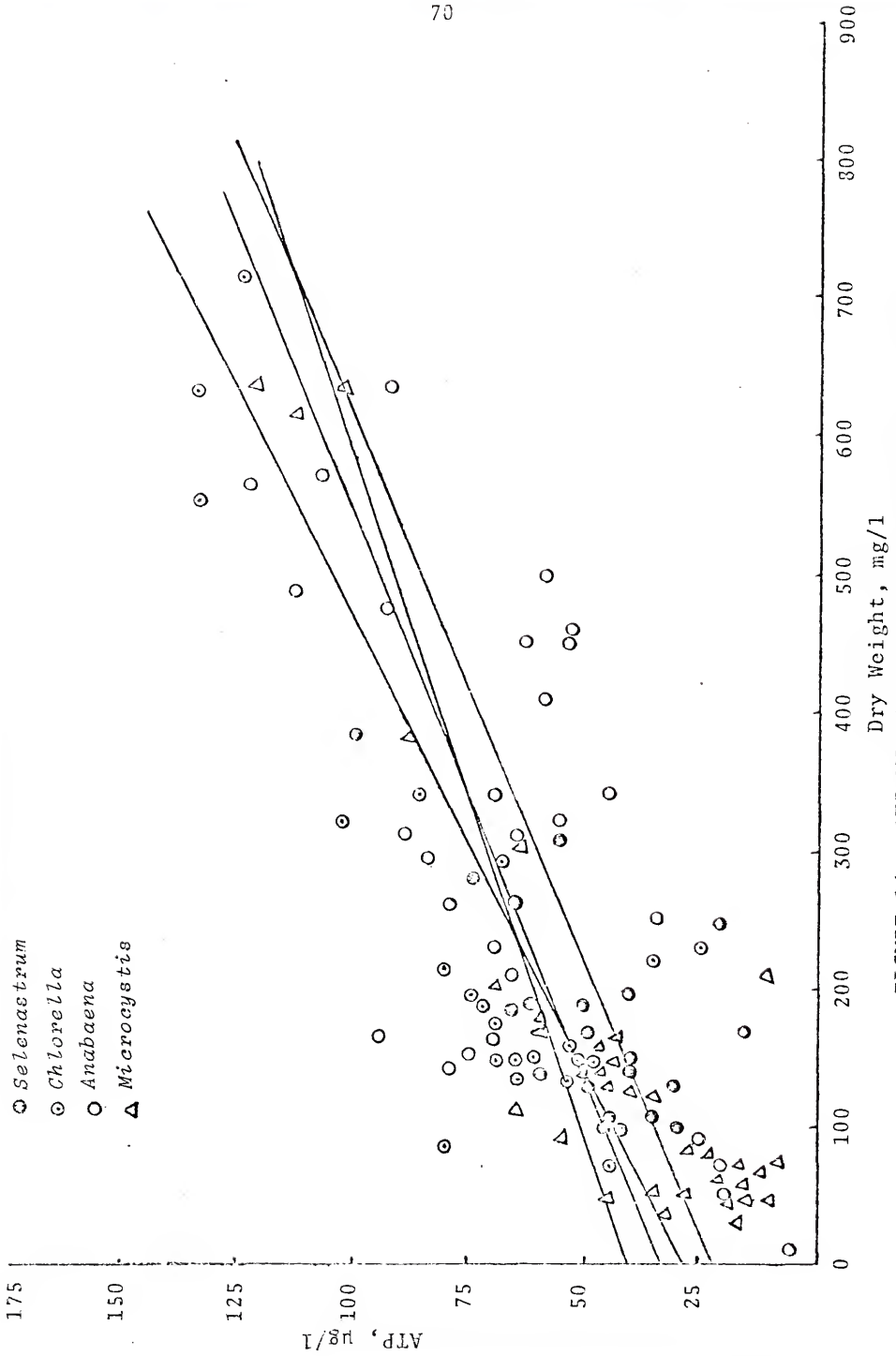


FIGURE 14. ATP VS. DRY WEIGHT OF UNIALGAL CULTURES

TABLE 10

ATP VS. DRY WEIGHT CORRELATION COEFFICIENTS  
FOR UNIALGAL CULTURES

Algae	Sample Correlation Coefficient, r	95% Confidence Interval, $\rho$	99% Confidence Interval, $\rho$
<i>Anabaena</i>	0.607	0.29 - 0.78	0.16 - 0.82
<i>Selenastrum</i>	0.244	-0.18 - 0.60	-0.33 - 0.68
<i>Microcystis</i>	0.718	0.41 - 0.87	0.27 - 0.91
<i>Chlorella</i>	0.860	0.65 - 0.94	0.56 - 0.85
All Algae	0.726	0.61 - 0.79	0.56 - 0.82

*Microcystis*, and *Chlorella* are relatively good ranging from 0.607 to 0.860, with all three significant at the 99 percent confidence level. Correlation at ATP with dry weight for *Selenastrum* is poor and not significant even at the 95 percent level. The poor correlation of *Selenastrum* probably resulted primarily from the small range of values measured. All of the algae together have a correlation coefficient of 0.726 which is significant at the 99 percent confidence level. In general, good correlation exists between ATP and dry weight in the unialgal cultures.

The ATP to dry weight measurements for Bivens Arm, Newnan's Lake, and Anderson-Cue Lake are shown in Figure 15. Correlation coefficients for the three lakes are all significant at the 99 percent confidence level (Table 11). Good correlation exists between ATP and dry weight for all three lakes, with Anderson-Cue Lake having the best correlation.

ATP to dry weight ratios for the algal cultures and the lakes were calculated and are presented in Table 12. All four algal cultures have relatively constant ATP to dry weight ratios ranging from 0.24 to 0.38  $\mu\text{g}/\text{mg}$ . However, the ATP to dry weight ratios for the three lakes vary from 0.12 to 0.96  $\mu\text{g}/\text{mg}$ . Bivens Arm and Newnan's Lake have a much lower average ATP to dry weight ratio than Anderson-Cue Lake. These low ratios are evidently the result of a high concentration of detrital matter which also explains the

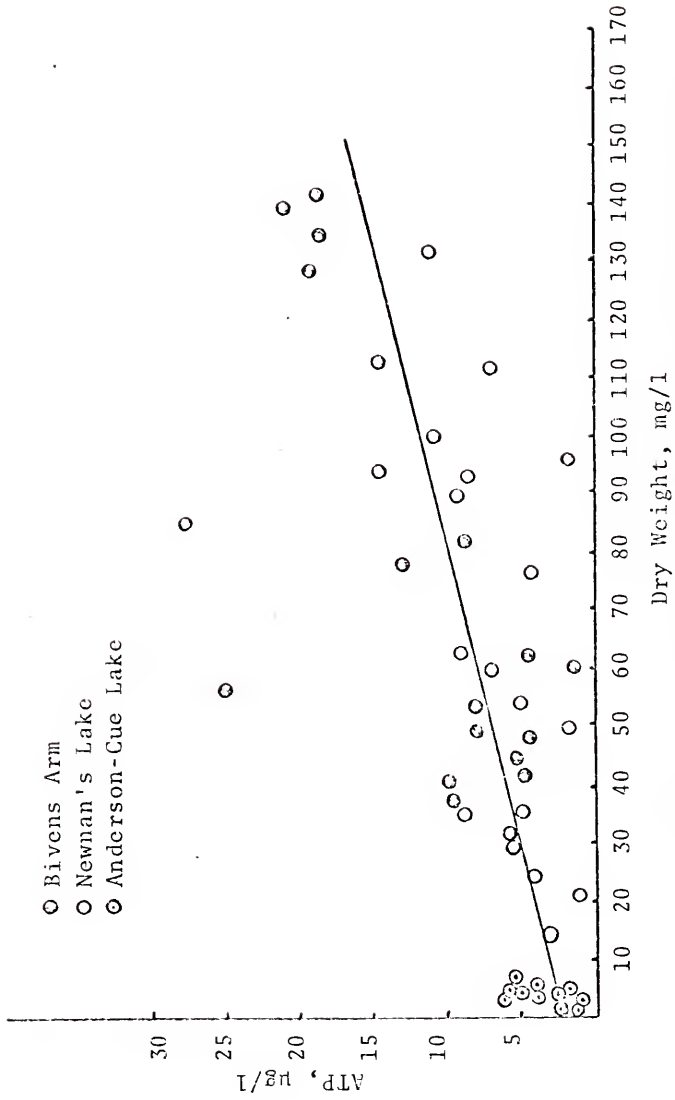


FIGURE 15. ATP VS. DRY WEIGHT FOR LAKE PHYTOPLANKTON

TABLE 11

ATP TO DRY WEIGHT CORRELATION COEFFICIENTS  
FOR LAKE PHYTOPLANKTON

<u>Lake</u>	<u>Correlation Coefficient, r</u>	<u>95% Confidence Interval, <math>\rho</math></u>	<u>99% Confidence Interval, <math>\rho</math></u>
Bivens Arm	0.649	0.20 - 0.82	0.15 - 0.87
Newnan's Lake	0.673	0.23 - 0.83	0.18 - 0.88
Anderson-Cue Lake	0.790	0.36 - 0.93	0.14 - 0.95

TABLE 12

ATP TO DRY WEIGHT RATIO'S FOR ALGAL CULTURES  
AND LAKE PHYTOPLANKTON

	<u>ATP/dry weight</u>
<i>Anabaena</i>	0.26
<i>Microcystis</i>	0.36
<i>Selenastrum</i>	0.24
<i>Chlorella</i>	0.38
Bivens Arm	0.13
Newnan's Lake	0.12
Anderson-Cue Lake	0.96

poorer correlation of ATP with dry weight in these lakes. Table 13 compares the turbidity of each lake with the measured ATP to dry weight ratio. Both Bivens Arm, and Newnan's Lake have large amounts of turbidity. Bivens Arm, for example, has an average turbidity of 10.2 JTU's compared to Anderson-Cue's average of 1.0 JTU.

TABLE 13

COMPARISON OF ATP/DRY WEIGHT RATIO WITH AVERAGE  
TURBIDITY OF EXPERIMENTAL LAKES

<u>Lake</u>	<u>ATP/Dry Weight</u>	<u>Average Turbidity<sup>1</sup> (JTU)</u>
Bivens Arm	0.13	10.2
Newnan's Lake	0.12	4.2
Anderson-Cue Lake	0.96	1.0

<sup>1</sup>Shannon, 1970

### ATP vs. Cell Number

Another common biomass measurement is the cell number. A dense culture of *Microcystis* was diluted into ten portions, incubated for 24-hours, and analyzed for ATP and cell number. Figure 16 shows that a linear correlation exists between ATP concentration and cell number. The ATP per cell for *Microcystis* ranged from  $5.0 \times 10^{-8}$  to  $6.25 \times 10^{-8}$   $\mu\text{g}/\text{cell}$ . It should be noted that cell number, because of differences in size between various algal species, would not give a useful correlation with ATP in mixed populations. These results are presented only to illustrate the good correlation between ATP and cell number in unialgal cultures.

### ATP vs. Absorbance

Absorbance is a common means of measuring relative algal densities in laboratory cultures (Joint Industry-Government Task Force on Eutrophication 1969). The absorbance of the four algal cultures was routinely measured along with ATP and the values are plotted vs. each other in Figure 17. Correlation coefficients for ATP vs. absorbance are given in Table 14. Except for *Microcystis* the correlation of ATP with absorbance is poor. A high correlation between dry weight and absorbance exists, indicating that scattering rather than absorbance is really being measured. It is surprising that better correlations were not obtained, especially for such dispersed algae as *Chlorella*. The reason lies partially in the small range of chlorophyll *a* values observed.

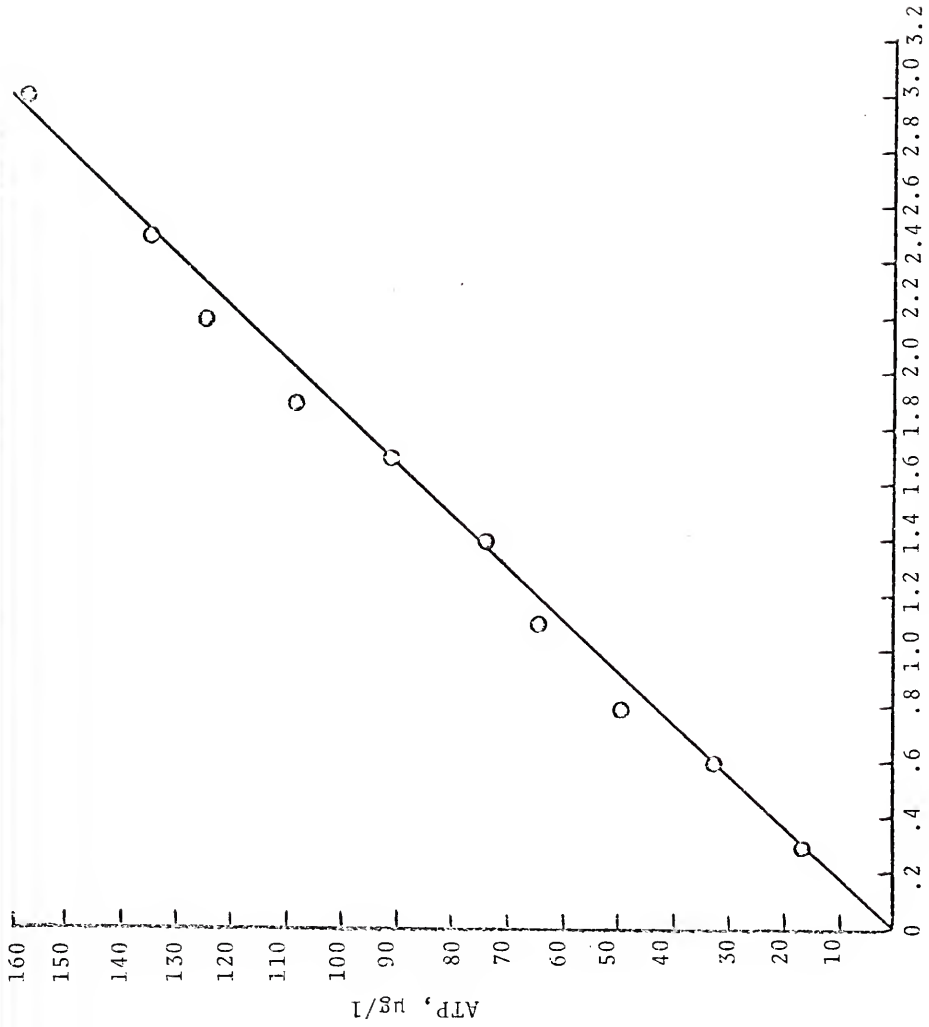


FIGURE 16. ATP VS. CELL COUNT FOR *MICROCYSTIS*



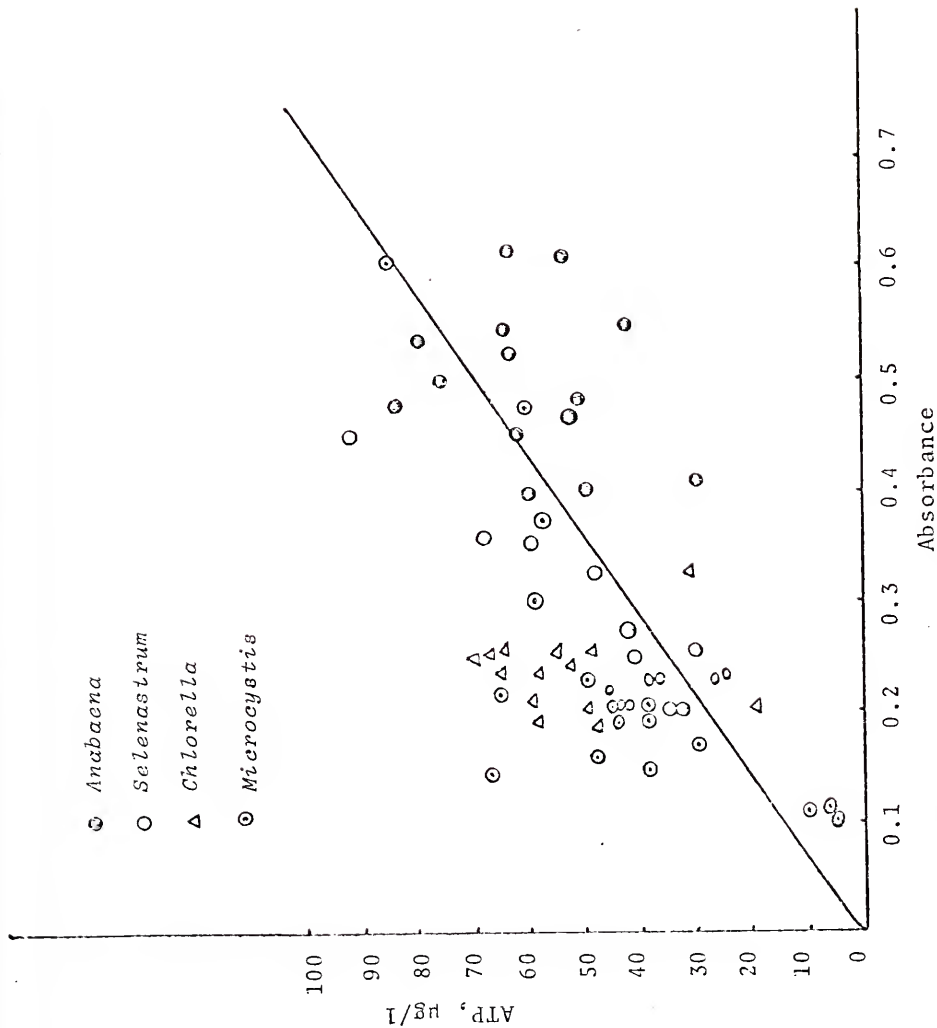


FIGURE 17. ATP VS. ABSORBANCE OF UNIALGAL CULTURES

TABLE 14

ATP TO ABSORBANCE CORRELATION COEFFICIENTS  
FOR UNIALGAL CULTURES

<u>Algae</u>	<u>Correlation Coefficient, r</u>	<u>95% Confidence Interval, p</u>	<u>99% Confidence Interval, p</u>
<i>Anabaena</i>	0.069	-0.42 - 0.48	-0.51 - 0.66
<i>Microcystis</i>	0.676	0.36 - 0.84	0.15 - 0.90
<i>Selenastrum</i>	0.236	-0.29 - 0.70	-0.47 - 0.75
<i>Chlorella</i>	0.152	-0.37 - 0.60	-0.53 - 0.71

### Zooplankton Study

In order to determine the effect of zooplankton ATP on the total measured ATP in natural plankton populations, the ATP of zooplankton from a local lake was measured. Zooplankton from Newnan's Lake was collected using a zooplankton net. The zooplankton consisted almost exclusively of Cyclops whose average dimensions were 1.4 mm by 0.3 mm. Using these average dimensions and assuming a specific gravity of 1.05, the ATP to dry weight ratio for zooplankton was calculated to be 0.1  $\mu\text{g}/\text{mg}$ . This value is slightly lower than the average value reported in this research for algal cultures and much lower than the average value reported for bacterial cultures (1.5  $\mu\text{g}/\text{mg}$ ). Based on the lower ATP to dry weight ratio of zooplankton and their relatively small number in most lakes, it seems that the ATP in zooplankton would not significantly add to the total ATP measured in a lake. Thus, these results indicate that ATP from zooplankton would not significantly affect the phytoplankton biomass estimated using ATP concentration.

### ATP - Response Results

The response of ATP to different environmental conditions was measured to determine the usefulness of ATP analysis as a rapid bioassay technique and to assess the stability of ATP in reference to its application to measurement of phytoplankton biomass. Laboratory algal

cultures and lake phytoplankton were subjected to varying light periods, different pH levels, and additions of nutrients and toxins.

#### Light-Dark Study

In order to be used as a reliable estimate of phytoplankton biomass, the cellular ATP concentration must not vary under light and dark conditions. To determine whether the cellular ATP concentration remained relatively constant under light and dark conditions the following experiment was performed. Aliquots of *Selenastrum* were placed alternately in the light and the dark for a period of ten hours and the ATP content was monitored. The response of ATP in *Selenastrum* cells to light and dark periods is shown in Figure 18. Fluctuations in the ATP content occurred whenever the algae were changed from either light to dark or from dark to light. Within 15 to 20 minutes after each transition the ATP concentration returned to the initial ATP level. Holm-Hansen (1970) obtained similar results working with marine algae, and found ATP concentrations to be maintained at fairly uniform levels in both light and dark. Holm-Hansen's experiment only covered a time span of 22 minutes while the above mentioned experiment covered a span of ten hours, indicating the long-term stability of ATP under light and dark conditions. The ATP pool in algae is probably regulated by the interaction of the photosynthetic

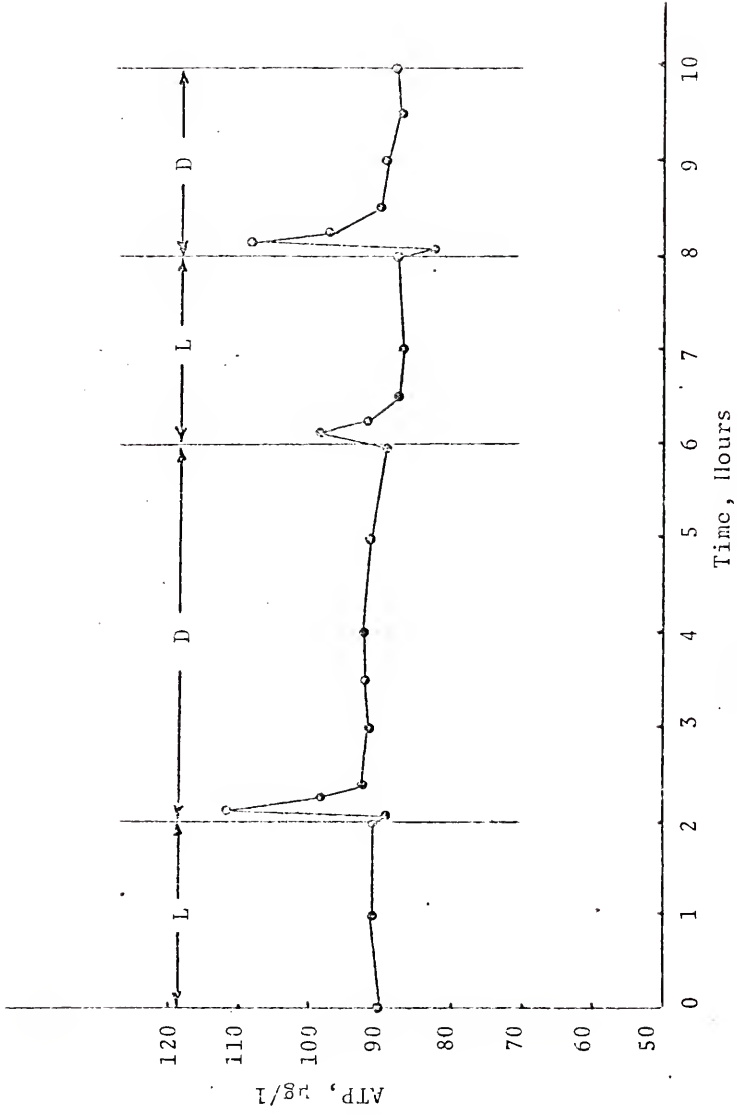
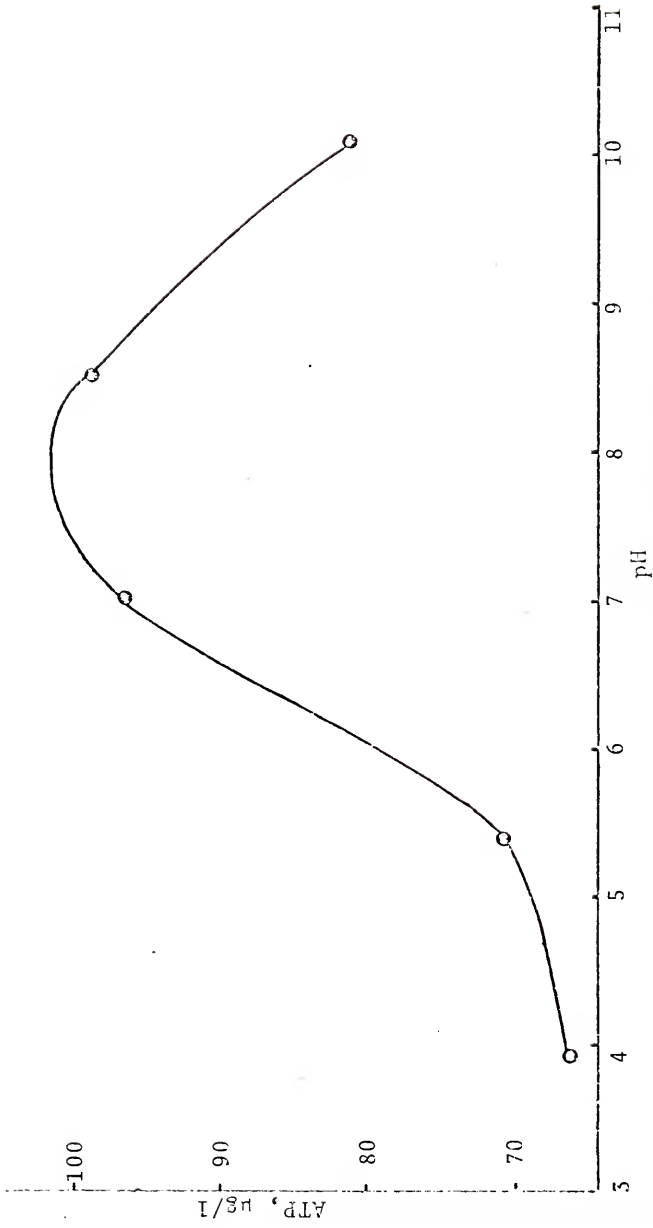


FIGURE 18. RESPONSE OF ATP TO PERIODS OF LIGHT AND DARK

and respiratory mechanisms of the cell. ATP occurs in trace amounts in the cell and has a rapid turnover rate. It is probable that ATP is primarily regulated by respiration rather than photosynthesis. In photosynthesis light energy is used to form ATP which is almost immediately used for simple sugars. These simple sugars enter the carbon cycle and eventually supply all the energy needs of the cell. However, the fluctuations in ATP levels observed at light to dark transitions indicate a definite interaction and dependence between photosynthesis and respiration.

#### pH Studies

The pH of natural waters is an important environmental variable. The biota, productivity and solubilities of nutrients and toxic substances are all somewhat dependent on the pH of a specific water. To use ATP as a measure of biomass, the relationship between ATP concentration and pH must be known. Thus, the response of cellular ATP level to different pH values was measured. Figure 19 shows the ATP concentration measured in *Selenastrum* cells after incubation for one hour at various pH levels. Maximum ATP concentration occurred in the pH range of 7.5 to 8.0, the normal growing range of the laboratory cultures. The ATP concentration was greatly reduced in the pH range of 4.0 to 5.0. Moderate reduction in ATP concentration occurred at a pH of 10.0.

FIGURE 19. EFFECT OF pH ON ATP CONTENT OF *SELENASTRUM* CELLS

The above experiment was repeated using lake water, in order to determine whether maximum ATP content always occurred within a neutral pH range such as pH 7 to 8, or whether the optimum ATP occurred at the normal ambient pH of the organisms. Figure 20 shows the ATP concentration measured in Anderson-Cue Lake water after incubation for one hour at various pH values. Maximum ATP concentration occurred at a pH of 4.6, the original pH of the lake water. The greatest reduction in ATP concentration occurred in the pH range of 9.0 to 11.0. A high pH lake water was tested in the same manner. Figure 21 shows the ATP concentration measured in Bivens Arm lake water after incubation for one hour at different pH levels. Maximum ATP concentration occurred at a pH of 9.0, the original pH of the lake water. Maximum reduction in ATP concentration occurred at a pH of 3.0.

It should be noted that the pH values reported are those of the incubation media or lakewater. It is probable that the cellular pH changed less than the media or lake water since the cell can control intracellular pH to some extent. The reduced ATP concentrations may indicate a reduction in viable cells or a shift in the energy balance of the cell as it attempts to maintain homeostasis. These results indicate that a shift from the organisms normal pH range causes a reduction in the ATP level, whether acidic, neutral, or basic.



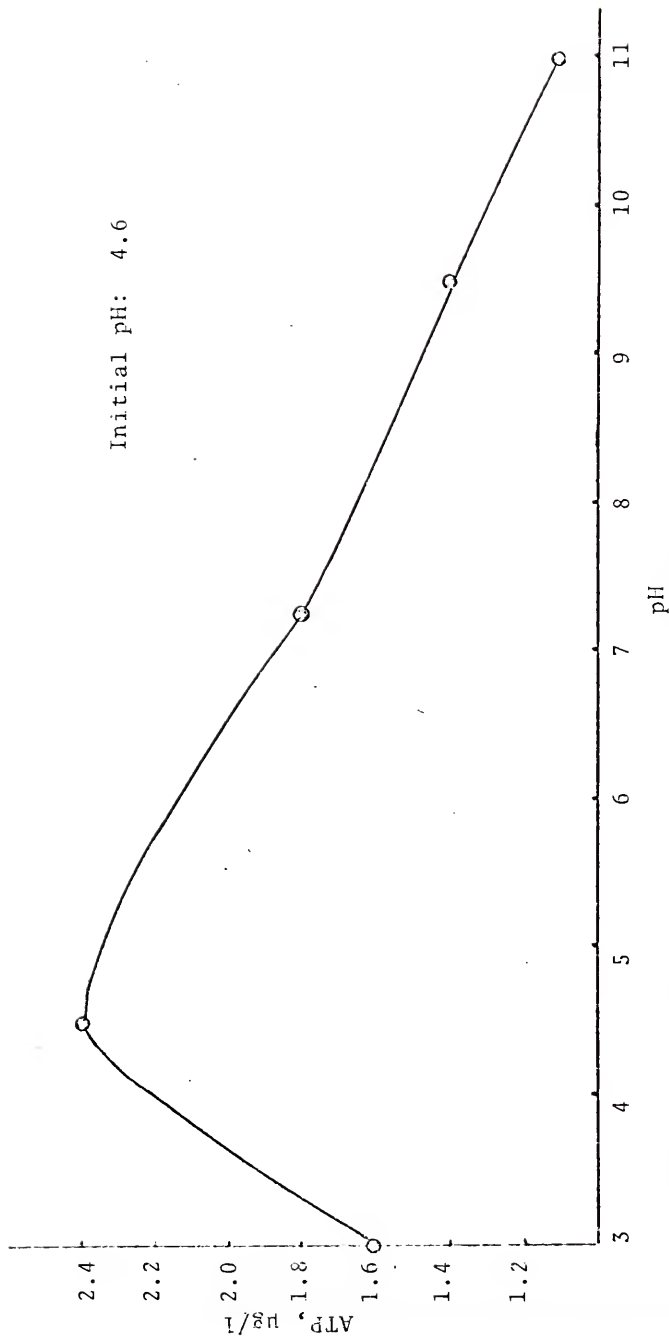


FIGURE 20. EFFECT OF pH ON ATP CONTENT OF ANDERSON-CUE LAKE WATER

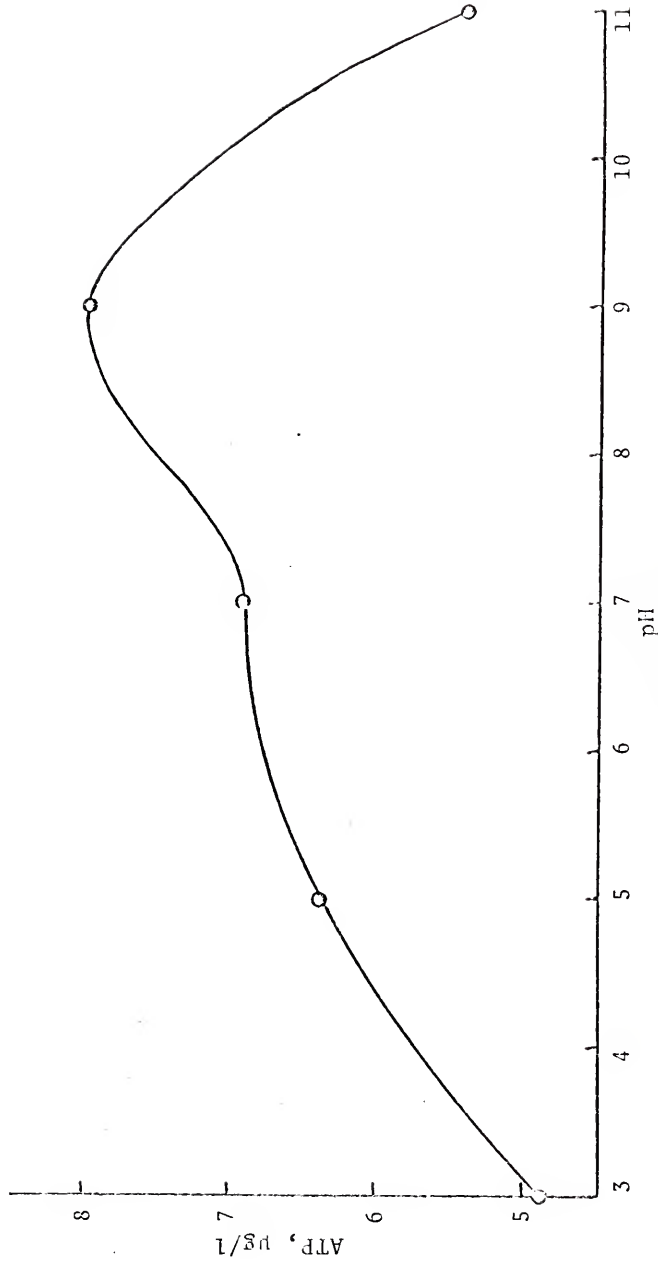


FIGURE 21. EFFECT OF pH ON ATP CONTENT OF BIVENS ARM LAKE WATER

### Toxicity Studies

One of the greatest problems today is the introduction of toxic materials into our waterways. In 1969 almost 70 percent of the reported fish kills were caused by the discharge of industrial wastes. Wastes containing concentrations of heavy metals may be toxic to aquatic organisms and severely affect the water community. The bioassay is an important tool in the investigation of toxic materials because the results of such a study indicate the degree of hazard to aquatic life of particular toxins. From bioassay studies interpretations and recommendations can be made concerning the level of discharge that can be tolerated by the receiving water.

Mercury compounds are highly toxic to aquatic communities. Mercuric chloride is used in disinfecting, preserving, tanning, electroplating and many other processes. Mercurio-organic compounds are used in herbicides, fungicides, and medical treatment. They have been used extensively to control slimes in paper mills. To study the effects of mercuric ion on phytoplankton, aliquots of *Selenastrum* were incubated with various concentrations of mercuric chloride and the response of ATP, carbon-14 uptake, chlorophyll *a*, and suspended solids was measured. Figure 22 shows the ATP content in *Selenastrum* samples after three hours of incubation in the presence of various concentrations of mercuric chloride. The ATP toxicity pattern in Figure 22 shows a

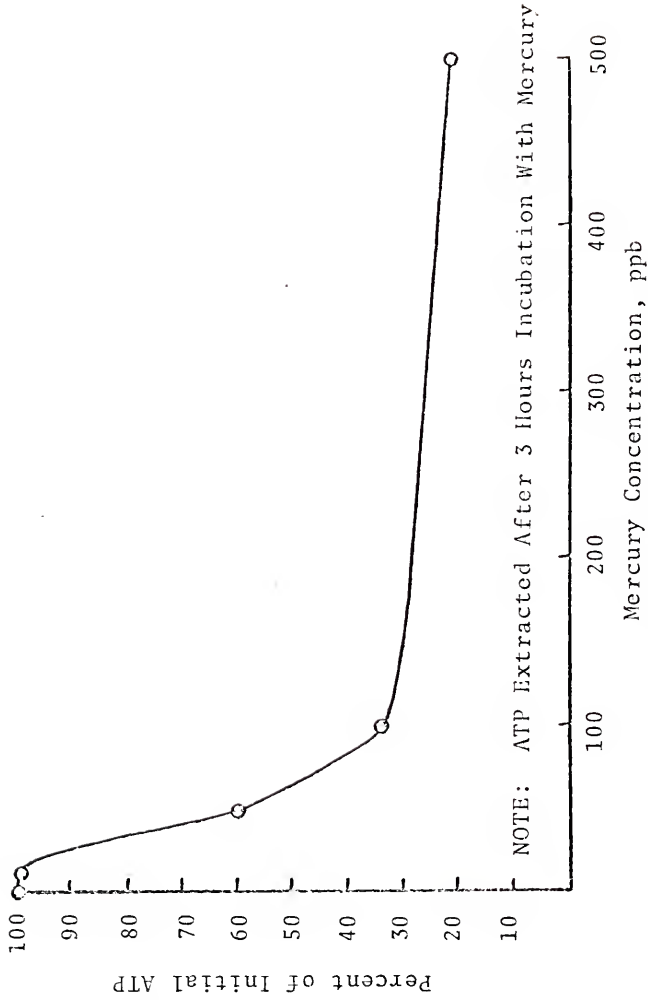


FIGURE 22. ATP POOL RESPONSE TO INCUBATION WITH VARIOUS CONCENTRATIONS OF MERCURY

rapid reduction of ATP for mercuric chloride concentrations of 50 ppb and 100 ppb, with a gradual further reduction for concentrations above 100 ppb. Table 15 shows the percent reduction in ATP, carbon-14 uptake, chlorophyll *a*, and suspended solids after one-hour incubation with 500 ppb of mercury.

TABLE 15  
PERCENT REDUCTION AFTER ONE-HOUR INCUBATION WITH  
MERCURIC CHLORIDE

<u>Parameter</u>	<u>Percent Reduction</u>
ATP	54
Carbon-14 Uptake	56
Chlorophyll <i>a</i>	20
Suspended Solids	10

From Table 15 it is apparent that ATP is a much more sensitive bioassay response parameter than chlorophyll *a* or suspended solids. However, ATP is about equal to or a little less sensitive than carbon-14 uptake as a response parameter.

The observed reduction in ATP may result from a reduction in the viable cells or it may result from a reduction in the cellular ATP pool. In an attempt to explain which mechanism was responsible for the reduction in ATP, a sample of *Chlorella* was incubated with 100 ppb of mercury for 80 minutes at which time half of the sample was centrifuged and the *Chlorella* cells were transferred to fresh PAAP media without mercury. The final volumes

of both samples were the same. As shown in Figure 25, the ATP content of the cells transferred to fresh media immediately increased from 104  $\mu\text{g}/\text{l}$  to 192  $\mu\text{g}/\text{l}$  while the ATP of the sample with mercury remained constant. After five minutes the ATP content of the transferred cells decreased to 125  $\mu\text{g}/\text{l}$ , a concentration slightly below the initial concentration of the culture before addition of mercury. These results indicate that most of the ATP reduction results from a reduction in the cellular ATP pool rather than a reduction in the viable cell population. The difference between the initial ATP concentration (135  $\mu\text{g}/\text{l}$ ) and the final concentration of the transferred cells (125  $\mu\text{g}/\text{l}$ ) indicates that a fraction of the algal cells were killed by the addition of mercury. Eighty minutes was selected as the time to transfer the poisoned *Chlorella* cells to fresh media because previous experiments showed that maximum ATP reduction occurred in the first 1 to 1 1/2 hours after addition of the poison. It is probable that as the incubation time increases the fraction of dead cells would increase. After two hours of incubation the ATP content of the transferred cells decreased slightly; this is probably a result of carry over of mercury in the centrifuged cells.

Another important toxic material is copper. Copper sulfate is used extensively in industry for tanning, electroplating, engraving, and pigment manufacture. It is also used to control undesirable plankton growth in reservoirs and lakes. In order to study the effect of copper on

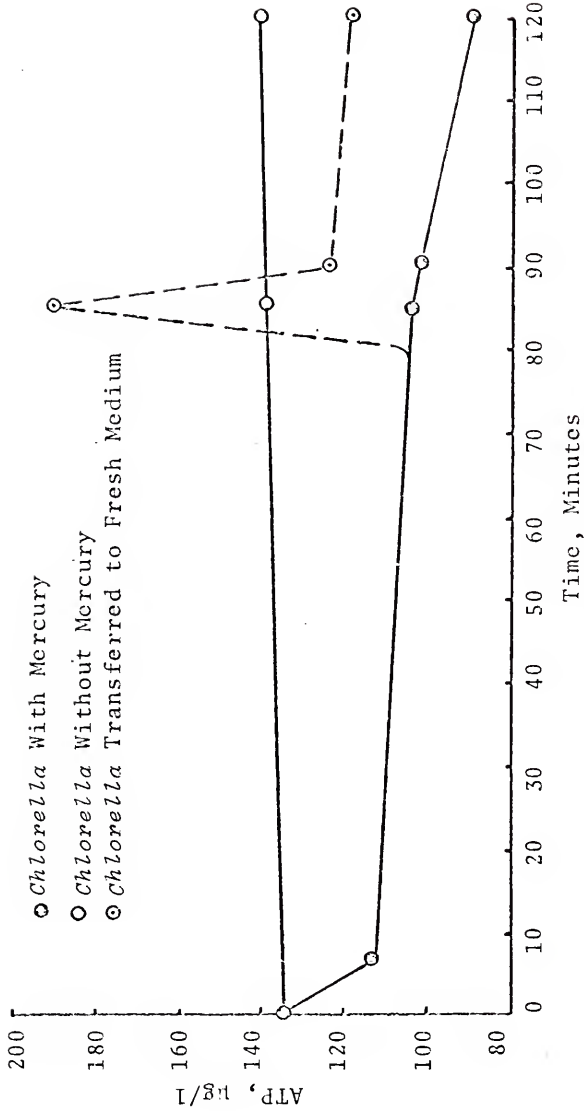


FIGURE 23. RESPONSE OF MERCURY-POISONED *CHLORELLA* AFTER TRANSFER TO FRESH MEDIUM

natural phytoplankton populations, Bivens Arm water was incubated with a concentration of 2 mg/l of copper sulfate. Bivens Arm, at the time of sampling, contained a very dense bloom of *Microcystis*. Figure 24 shows the response of ATP, chlorophyll  $\alpha$ , suspended solids, and COD to incubation with copper sulfate. In the first three hours of incubation the ATP content dropped by 45 percent while carbon-14 uptake, suspended solids, chlorophyll  $\alpha$ , and COD dropped 40 percent, 22 percent, 29 percent, and 21 percent respectively. The ATP concentration decreased much more rapidly than all of the other parameters with the exception of carbon-14 uptake. Both ATP and carbon-14 uptake decreased rapidly after addition of copper sulfate. The rapidity of the decrease in ATP and its similarity to the decrease in carbon-14 uptake indicates that a decrease in cellular ATP occurs initially rather than a decrease in living cells. These results indicate that ATP analysis is a sensitive response method that can be useful in routine toxicity bioassays.

#### Nutrient Studies

Besides bioassays for determining the effect of toxic substances on aquatic communities, there is a need for better limiting nutrient. Considerable controversy exists today as to whether phosphorus, nitrogen, or carbon is the primary limiting nutrient. These three nutrients were used to determine the usefulness of ATP analysis as a rapid bioassay method.



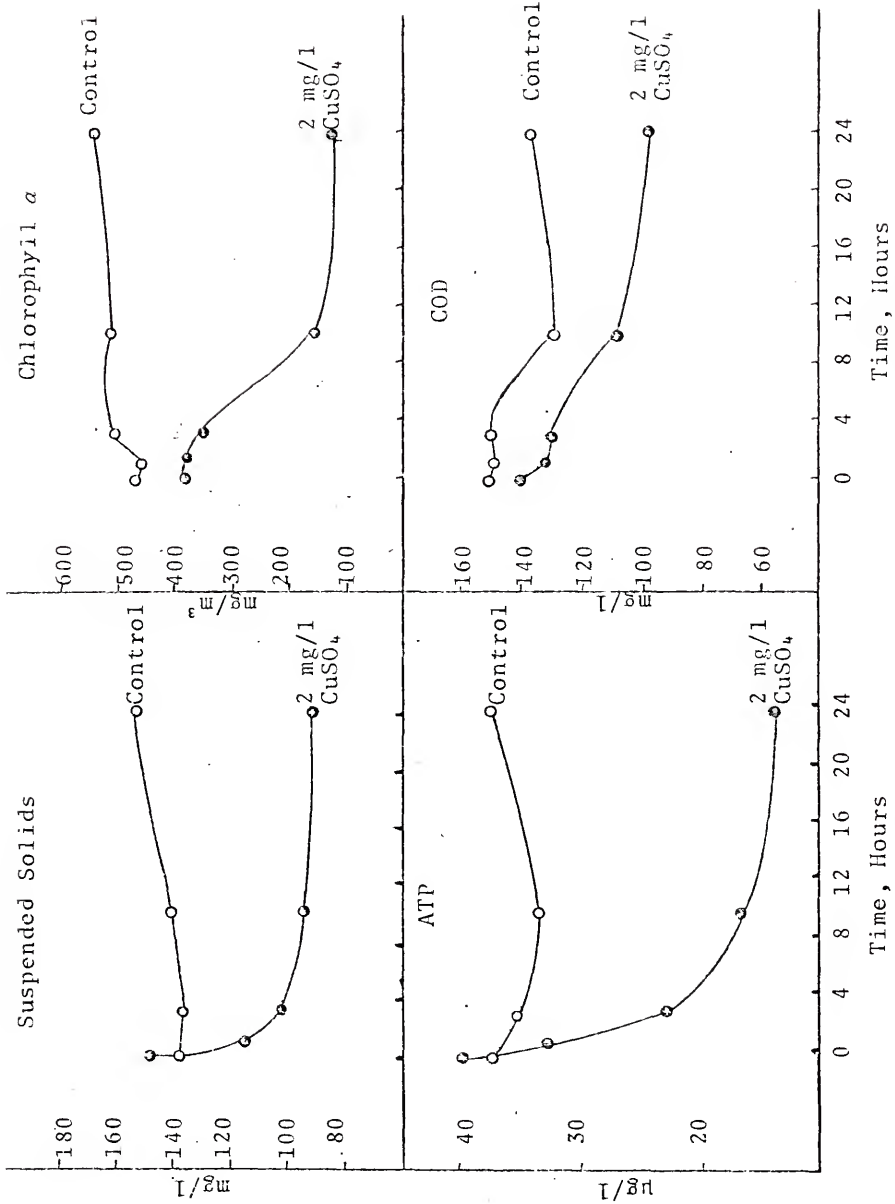
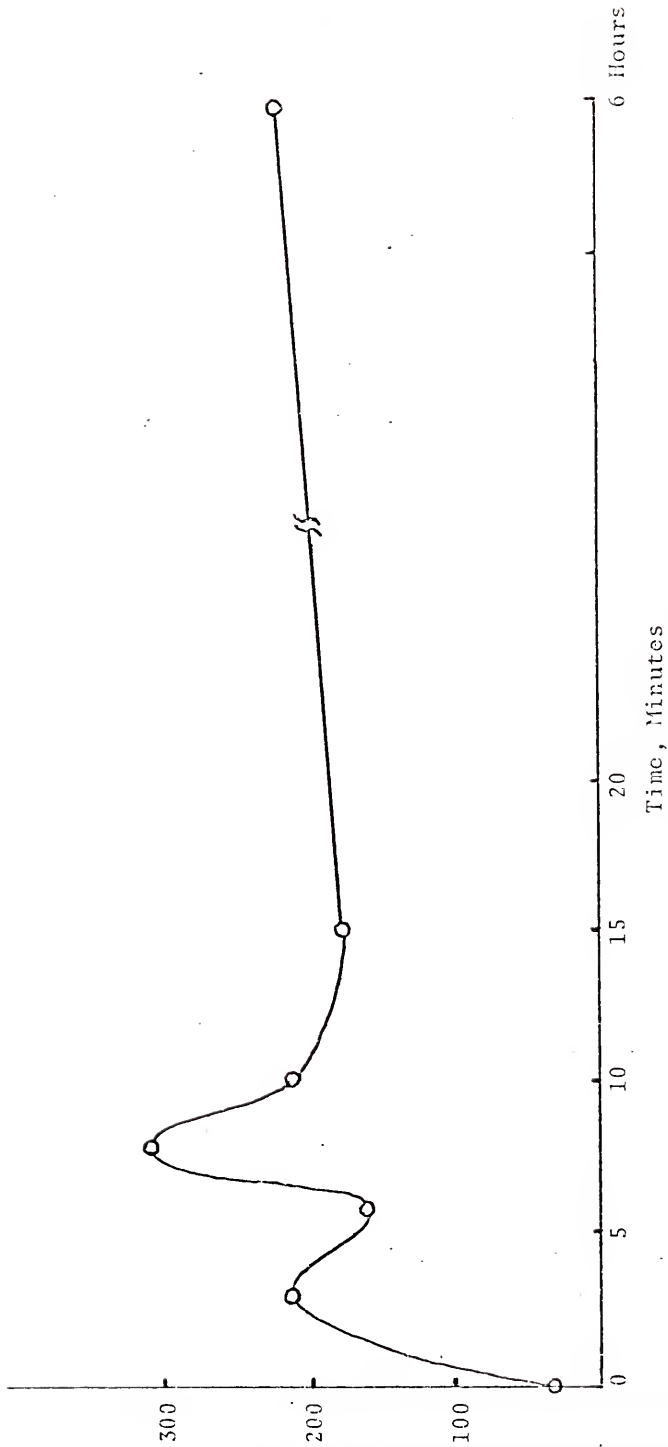


FIGURE 24. RESPONSE OF BIOMASS PARAMETERS IN BIVENS ARM TO ADDITION OF COPPER

The effect of adding carbon dioxide to carbon-limited algae was determined by bubbling carbon dioxide gas into a batch culture of *Anabaena* and measuring the response in the ATP concentration. The *Anabaena* culture had been grown for three days on PAAP medium. Initially the pH of the culture was 7.6 but each day it increased until it reached 10.5 on the day of the experiment. At this pH the culture was carbon-limited, very dense, and a dull green. Past experiments indicated that carbon-limited *Anabaena* become dull green, but change to a bright green color approximately one day after adding carbon-dioxide. In this experiment the carbon-dioxide was bubbled into the culture for five seconds. The response of ATP to addition of carbon dioxide was dramatic as shown in Figure 25. Cellular ATP increased from 154  $\mu\text{g}/\text{l}$  to 312  $\mu\text{g}/\text{l}$  in three minutes. The ATP content was measured for 12 minutes and exhibited a tendency to rapidly increase and decrease. After six hours the ATP content still remained relatively constant at 312  $\mu\text{g}/\text{l}$ . The significant increase in ATP concentration and its stability at the higher concentration may also indicate that the original sample at pH 10.5 may have been inhibited by the high pH. The higher ATP concentrations may be a result of increased cellular activity, utilizing the available carbon. Immediately after carbon dioxide addition the pH dropped to 5.8, but after six hours the pH increased to 8.6, indicating that the carbon was being utilized by the algae.

FIGURE 25. ATP RESPONSE IN *A. BAEBANA* TO ADDITION OF CARBON DIOXIDE

Another experiment on carbon dioxide addition was performed with a natural population from Bivens Arm. The Bivens Arm sample had a mixed algal population with a low density. The initial pH of the sample was 9.0. Figure 26 shows the ATP response in Bivens Arm water to which carbon dioxide was added. Like the algal culture, an initial increase in ATP occurred. However, unlike the algal culture, the ATP concentration decreased shortly thereafter and continued to decrease, indicating inhibition or die-off of the phytoplankton. After three minutes the pH dropped to 6.7 and after 23 minutes it dropped to 5.7. The decrease in ATP is probably a result of the lowered pH of the lake water. An increase in the pH to 7.6 was observed after three hours, an indication that the lake water was recovering from the carbon dioxide addition.

The response of ATP in phosphorus-deficient cultures to the addition of phosphorus was also measured. Batch cultures of *Anabaena*, *Microcystis*, *Selenastrum*, and *Chlorella* were grown in duplicate for a period of five days. One culture of each algal species contained regular PAAP medium with 0.62 mg/l while the other culture contained a modified PAAP medium without phosphorus. All cultures were relatively dense having an absorbance ranging from 0.19 to 0.57. On the fifth day 1.24 mg/l of phosphorus was added to each culture and the ATP response measured. Table 16 gives the percent increases of ATP after addition of phosphorus. All

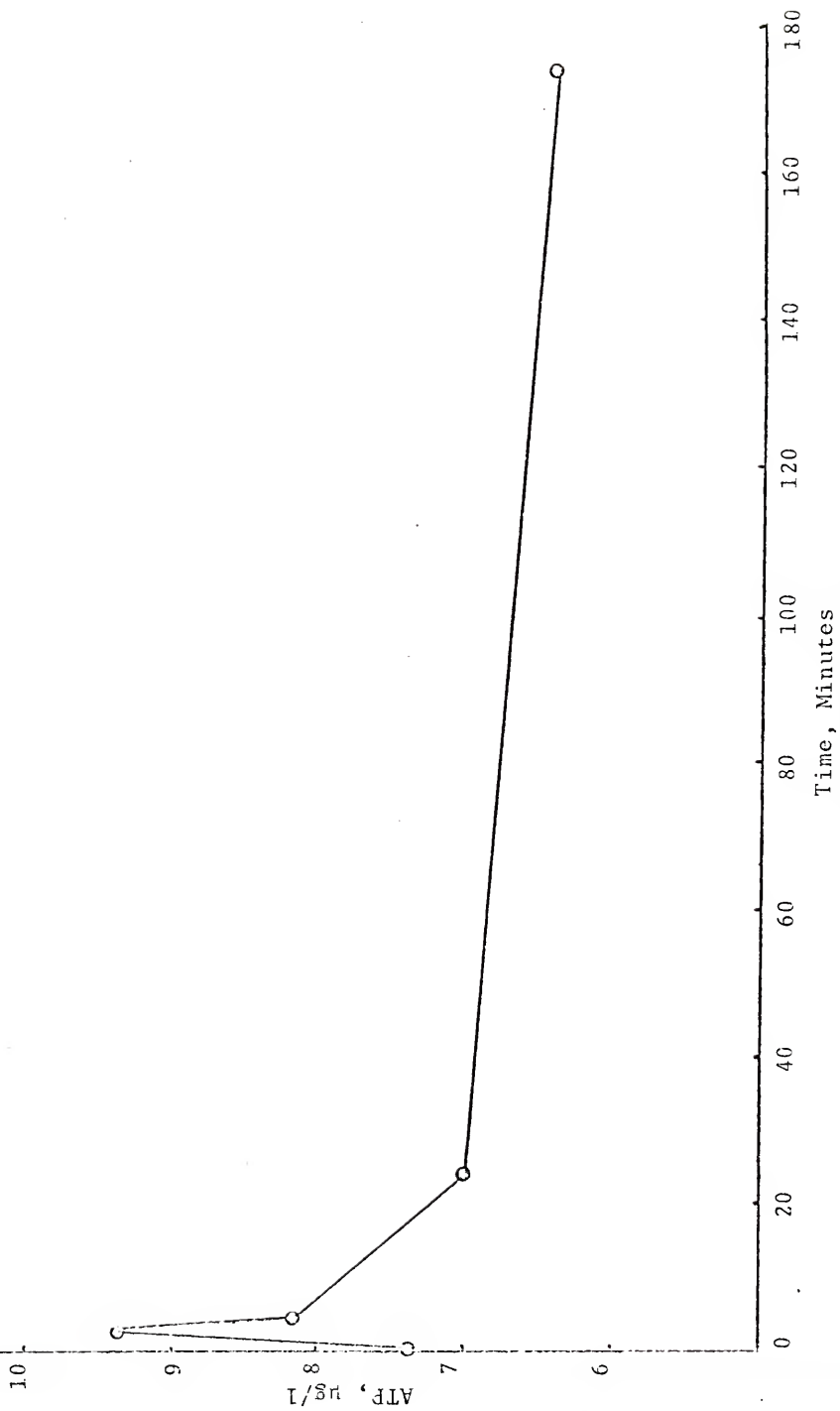


FIGURE 26. ATP RESPONSE OF BIVENS ARM TO ADDITION OF CARBON DIOXIDE

TABLE 16

## RESPONSE OF CELLULAR ATP TO ADDITION OF PHOSPHATE

Algal Culture	Phosphate Concentration		Percent Increase in ATP <sup>1</sup>
	Initial	Final	
<i>Anabaena</i> <sup>2</sup>	.52	.12	10
<i>Anabaena</i> <sup>3</sup>	.12	.07	80
<i>Selenastrum</i> <sup>2</sup>	.28	.14	13
<i>Selenastrum</i> <sup>3</sup>	.05	.03	36
<i>Chlorella</i> <sup>2</sup>	.58	.12	5
<i>Chlorella</i> <sup>3</sup>	.02	.001	23
<i>Microcystis</i> <sup>2</sup>	.48	.03	12
<i>Microcystis</i> <sup>3</sup>	.16	.01	38

<sup>1</sup> Measured approximately 80 minutes after phosphorus addition

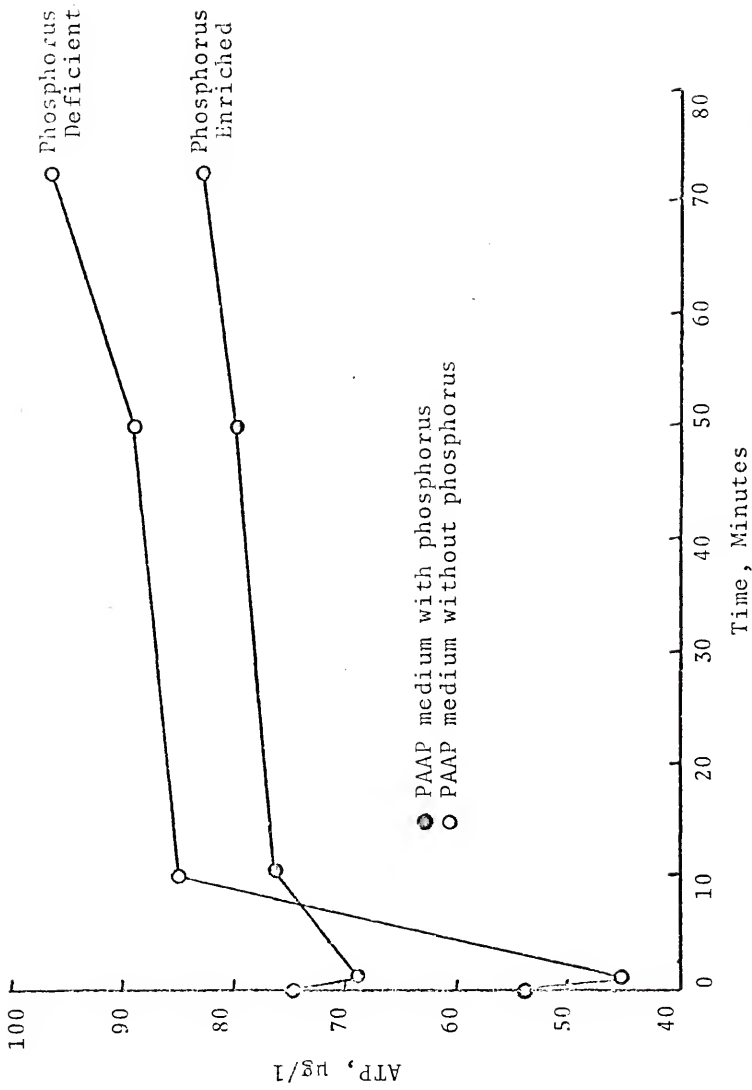
<sup>2</sup> Standard PAAP medium with 0.62 mg/l phosphorus

<sup>3</sup> Modified PAAP medium without phosphorus

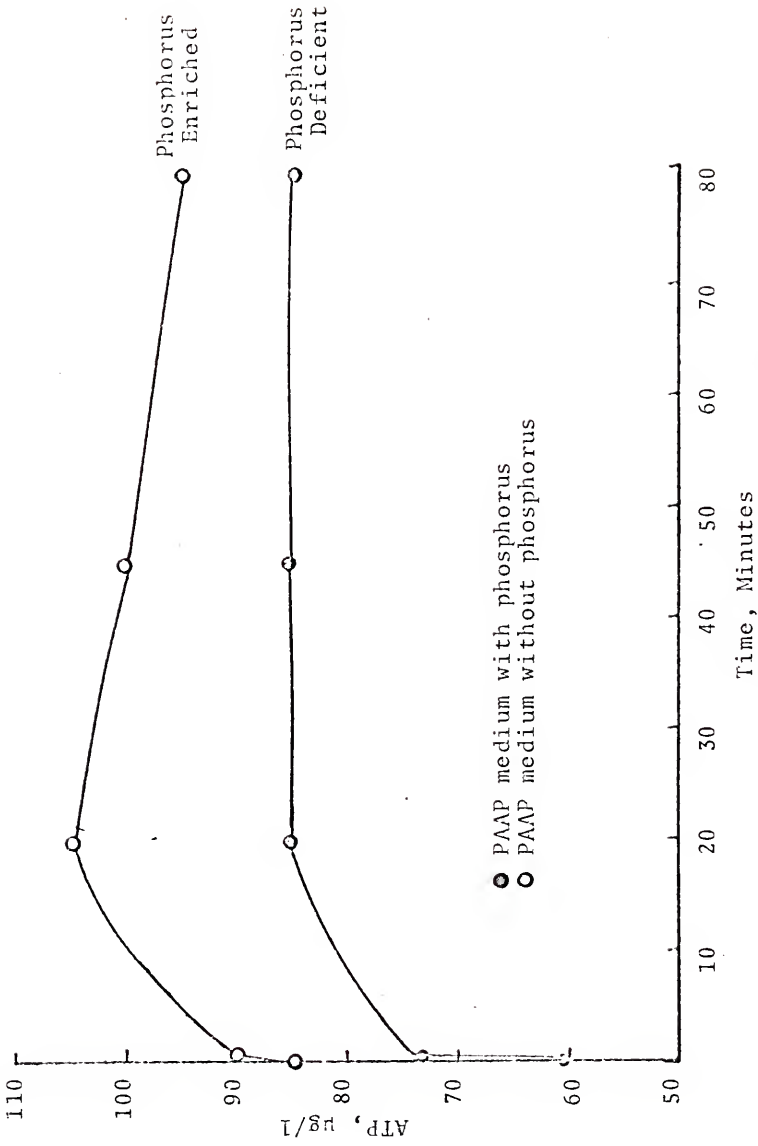
phosphorus-deficient cultures exhibited a greater ATP increase than the cultures with phosphorus. *Anabaena* showed the greatest difference in ATP response with the phosphorus deficient culture increasing eight times as much as the control culture. It is interesting to note that although both cultures of *Microcystis* had very low final concentrations of phosphorus, the ATP of the phosphorus-deficient culture increased approximately three times as much as the phosphorus-enriched culture. This would indicate that even though the phosphorus concentration of the phosphorus-enriched culture was low, the algal cells were not phosphorus limited, that is, at least not to the same extent.

The response patterns for each of the species are different as shown in Figures 27 to 30. *Anabaena* (Figure 27) exhibited an initial drop in ATP followed by an increase while *Microcystis* and *Selenastrum* (Figures 28 and 29 respectively) showed a rapid initial increase. *Chlorella* (Figure 30) displayed a gradual increase in ATP. All of the species, however, responded essentially in the same manner, that is, each exhibited an increase in ATP concentration.

To study the ATP response in nitrogen deficient cultures, *Anabaena* and *Selenastrum* were grown for five days in duplicate cultures. One culture contained PAAP medium with 14.0 mg/l of nitrogen as nitrate while the other contained a modified PAAP medium without nitrogen. On the

FIGURE 27. RESPONSE OF ATP IN *ANABAENA* TO ADDITION OF PHOSPHORUS



FIGURE 28. RESPONSE OF ATP IN *MICROCYSTIS* TO ADDITION OF PHOSPHORUS

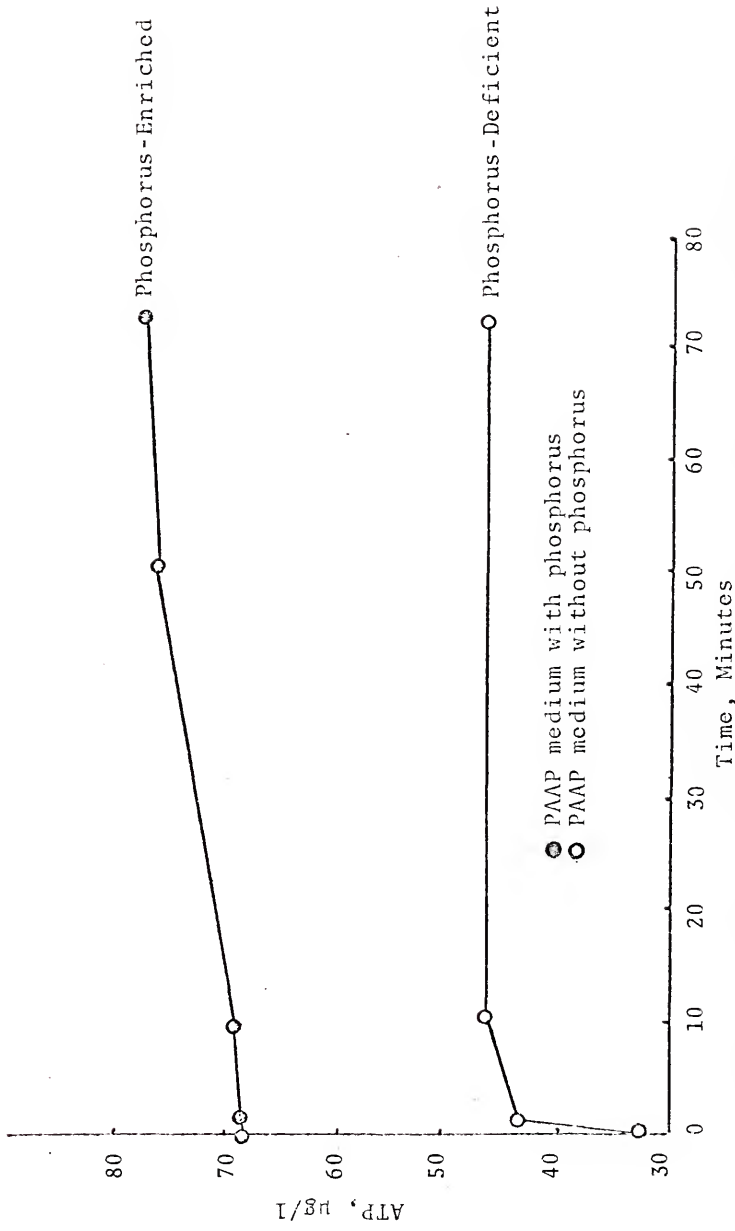


FIGURE 29. RESPONSE OF ATP IN *SELENASTRUM* TO ADDITION OF PHOSPHORUS

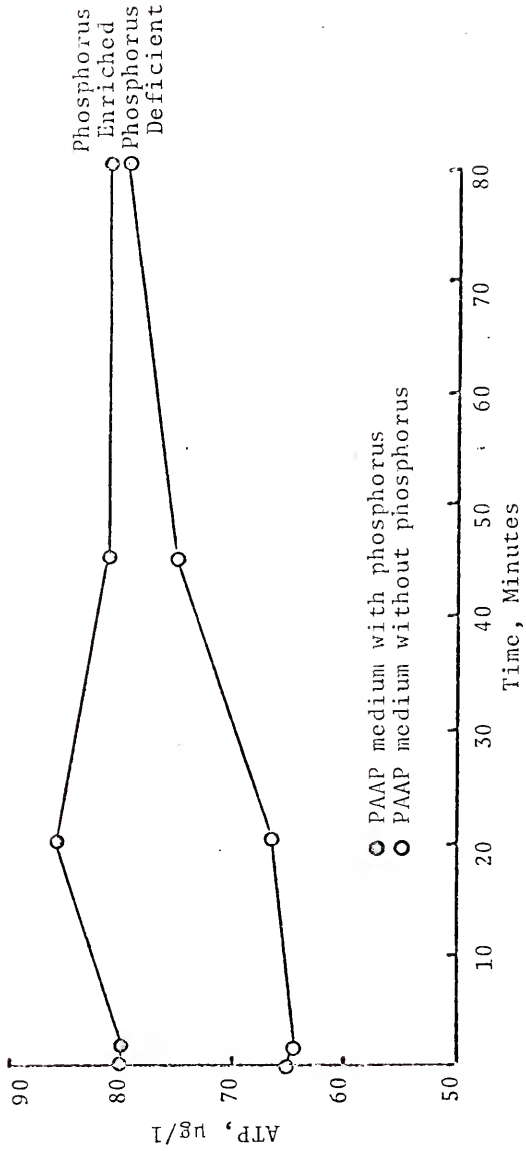


FIGURE 30. RESPONSE OF ATP IN *CHLORELLA* TO ADDITION OF PHOSPHORUS

fifth day, 23 mg/l of nitrogen as nitrate was added to each culture. Percent increases in ATP concentration after addition are given in Table 17. Like the phosphorus experiment, *Anabaena* exhibited the greatest difference in ATP response which is surprising because *Anabaena* is a nitrogen fixer while *Selenastrum* is not. The ATP response in nitrogen-deficient *Anabaena* was 46 times as great as that of the control. The ATP increase in the nitrogen-deficient *Selenastrum* was only twice that of the increase in the nitrogen-enriched culture. The response of the nitrogen-deficient *Anabaena* showed a gradual increase in ATP while the *Anabaena* control showed a rapid initial increase followed by a gradual decrease to approximately the original concentration (Figure 31). Both the nitrogen-deficient *Selenastrum* and the control responded rapidly as shown in Figure 32. Essentially, the ATP response to the addition of nitrogen was similar to that of phosphorus addition in that an increase in ATP concentration always occurred.

The percent increases in ATP concentrations reported here were calculated using the ATP values measured 70 to 80 minutes after the addition of nitrogen. This time interval was selected because the ATP content of both the nitrogen-deficient and the nitrogen-enriched cultures seemed to level off by this time. Different percent changes in ATP could be obtained simply by choosing a shorter or longer interval.

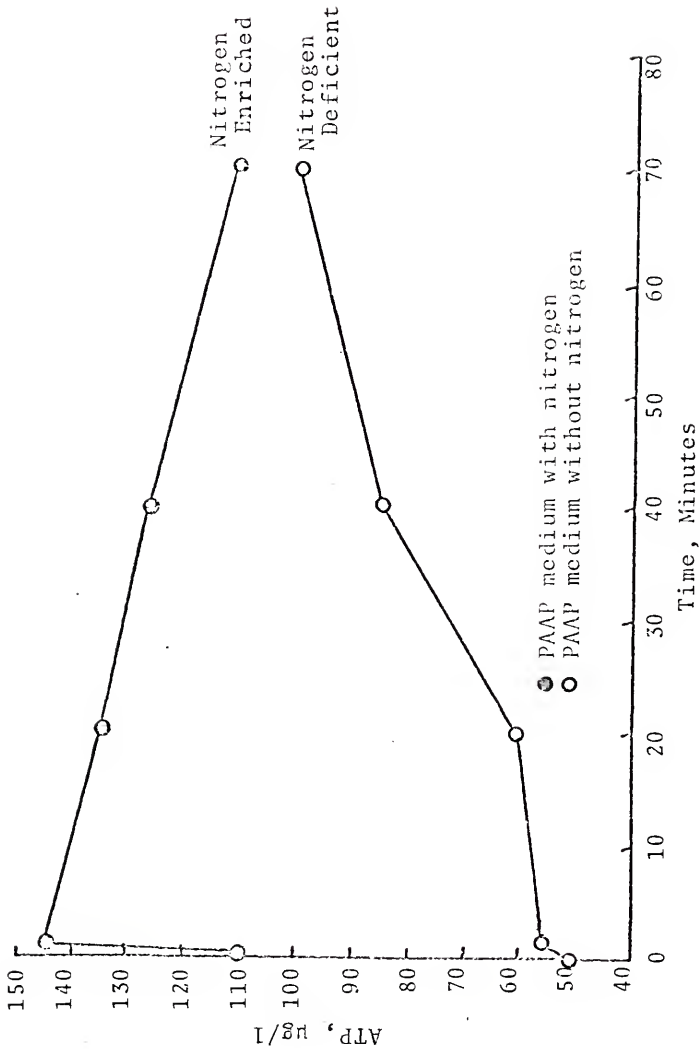


FIGURE 31. RESPONSE OF ATP IN *ANABAEHA* TO ADDITION OF NITROGEN

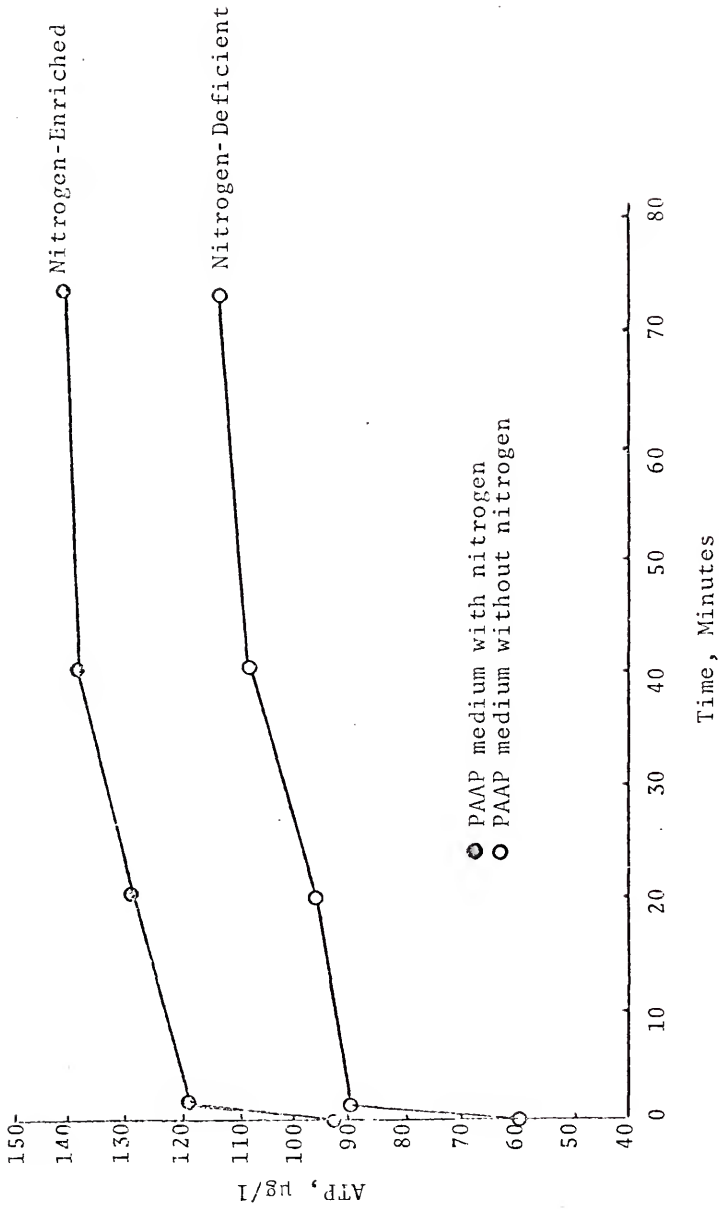


FIGURE 32. RESPONSE OF ATP IN *SELENASTRUM* TO ADDITION OF NITROGEN

TABLE 17

## RESPONSE OF CELLULAR ATP TO NITRATE ADDITION

<u>Algal Culture</u>	<u>Nitrate Concentration</u>		<u>Percent Increase in ATP<sup>1</sup></u>
	<u>Initial</u>	<u>Final</u>	
<i>Anabaena</i> <sup>2</sup>	6.70	1.4	2
<i>Anabaena</i> <sup>3</sup>	0.12	0.003	92
<i>Selenastrum</i> <sup>2</sup>	7.50	5.6	49
<i>Selenastrum</i> <sup>3</sup>	0.40	0.1	80

<sup>1</sup>Measured approximately 70 minutes after nitrogen addition

<sup>2</sup>Standard PAAP medium with 14.0 mg/l nitrogen

<sup>3</sup>Modified PAAP medium without nitrogen

For example, if the initial increase was used, the same response would be shown by both cultures in several instances.

This criticism is valid for the phosphorus experiments as well. The percent increases shown in Tables 16 and 17 for phosphorus and nitrogen respectively are presented merely to illustrate the differences observed in nutrient-deficient and nutrient-enriched cultures. The values presented in these tables indicate that ATP could be used as a rapid nutrient bioassay parameter. However, to be of practical value, the ATP response to nutrient addition should be measured over a period of a few hours and plotted against time. Plots of ATP vs. time would be a qualitative but relatively good tool in nutrient bioassays.



## CHAPTER VII. DISCUSSION

The experimental results presented in Chapter VII were briefly discussed as they were reported. In this chapter these results will be discussed in further detail and compared with work reported by other investigators.

Correlation of ATP with chlorophyll *a* for both unialgal samples and lake phytoplankton was good. The scatter of values around the linear regression line may be a result of changes in the cellular ATP pool but it is also likely that it may result from changes in cellular chlorophyll content. The lake samples were collected at different times of the day, during different seasons, and under different atmospheric conditions. Thus, the cellular chlorophyll may have varied diurnally. Ryther and Yentsch (1957) and Odum *et al* (1958) reported diurnal variations in chlorophyll concentrations. The chlorophyll concentrations may also have varied with nutritional state as reported by Fogg (1965) and others. Atmospheric conditions, such as a clear sky or an overcast sky, would also affect the chlorophyll concentration.

Holm-Hansen (1969) found good correlation between ATP and chlorophyll. He found that both ATP and chlorophyll gave good estimates of the cellular carbon in phytoplankton in seawater.

The ratios of ATP to chlorophyll *a* ranged from 0.09 to 0.35 for unialgal cultures and from 0.09 to 0.69 for lake phytoplankton. These values are in excellent agreement with those reported by Holm-Hansen (1969, 1970). His values ranged from 0.04 to 0.60 for five species of marine algae and one species of freshwater algae, and from 0.22 to 0.36 for seawater phytoplankton.

Good correlation of ATP with dry weight of unialgal cultures and lake phytoplankton was also observed. The ATP to dry weight ratios ranged from 0.24 to 0.38  $\mu\text{g}/\text{mg}$  for unialgal cultures and from 0.12 to 0.96  $\mu\text{g}/\text{mg}$  for lake phytoplankton. Holm-Hansen (1966) reported average ATP to dry weight values ranging from 0.15 to 1.6  $\mu\text{g}/\text{mg}$  dry weight for various marine algae (Table 18).

TABLE 18  
ATP TO DRY WEIGHT RATIO'S OF VARIOUS MARINE ALGAE  
(Holm-Hansen 1966)

<u>Organism</u>	<u>Average ATP/Dry Weight (<math>\mu\text{g}/\text{mg}</math>)</u>
<i>Skeletonema costatum</i>	0.31
<i>Amphidinium carteri</i>	0.15
<i>Dunaliella tertiolecta</i>	1.5
<i>Syracosphaera elongata</i>	0.03
<i>Monochrysis lutheri</i>	0.28
<i>Cyclotella nana</i>	0.84
<i>Ditylum brightwellii</i>	1.2
<i>Rhizosolenia sp.</i>	1.6

Lee *et al* (1971) reported ATP to dry weight values of *Selenastrum capricornutum* ranging from 1.4 to 3.4  $\mu\text{g}/\text{mg}$  dry weight. These values are much higher than those reported for this research. However, the authors note that they did not calculate the ATP to dry weight ratio after four days of incubation since the biomass after four days was known to consist of dead as well as living cell material. Table 19 lists the ATP to dry weight ratios reported by Lee *et al* (1971). Three cultures were grown with three different concentrations of phosphorus: 0.62 mg/l, 0.062 mg/l, and 3.1 mg/l.

TABLE 19

ATP TO DRY WEIGHT RATIO'S OF *SELENASTRUM* (Lee *et al* 1971)

<u>Phosphorus Content (mg/l)</u>	<u>Incubation Time</u>	<u>ATP/Dry Weight (<math>\mu\text{g}/\text{mg}</math>)</u>
0.62	4	3.1
	8	2.4
	16	0.7
	32	0.57
0.062	4	1.4
	8	0.47
3.1	4	3.4

The results in Table 19 indicate that the ATP to dry weight ratio varies with nutritional state. It is questionable whether the decrease in the ATP to dry weight ratio results primarily from a build-up of detrital material or from a physiological change due to a decrease in nutrition.

Correlation of ATP with cell number was also good. Holm-Hansen (1970) reported excellent correlation between ATP/cell and organic carbon/cell. Lee *et al* (1971) reported ATP contents ranging from  $3.9 \times 10^{-8}$  to  $6.5 \times 10^{-8}$   $\mu\text{g}/\text{cell}$  for *Selenastrum capricornutum* grown with 0.62 mg/l phosphorus and  $1.9 \times 10^{-8}$  to  $2.1 \times 10^{-8}$   $\mu\text{g}/\text{cell}$  for *Selenastrum capricornutum* grown with 0.062 mg/l phosphorus.

Cellular regulation of ATP was demonstrated by subjecting algal cells to alternating periods of light and dark. The ATP concentration fluctuated but always returned to its initial level. Santarius and Heber (1965) showed that ATP increased rapidly when an isolated chloroplast was exposed to light. When placed in the dark, the ATP decreased rapidly. They concluded that the ATP to ADP ratio controlled respiration in plants. When photosynthesis increases under high light intensity, it is believed that photorespiration occurs to stabilize the cell (Lehninger 1970). According to Lehninger, photorespiration does not occur in the mitochondria like regular respiration; it seems to be a wasteful way of regulating cells which are photosynthesizing too fast. Thus, evidence indicates that ATP is regulated by an intricate balance between photosynthesis, photorespiration, and respiration.

The response of ATP to various incubation pH levels indicates that cellular ATP decreases when an organism is shifted to a pH level different than its normal level.

Patterson (1970) reported similar results for activated sludge. He found the optimum ATP pool size occurred in the pH range of 7.5 to 8.0, the normal operating range of the activated sludge process.

Addition of toxic chemicals to unialgal cultures and lake phytoplankton caused a rapid decrease in ATP content. Patterson (1970) reported rapid decreases in ATP of activated sludge incubated with various toxins including nickel, chromium, chromate, mercury, copper, and cyanide.

Nutrient-deficient cultures of algae exhibited an increase in ATP content when nutrients were added. Holm-Hansen (1970) reported similar results for marine algae. He found that the ATP content increased when nitrogen was added to nitrogen-deficient cultures of *Monochrysis lutheri* and *Skeletonema costatum*. He also reported an increase in ATP content after addition of phosphorus to phosphorus-deficient cultures of *Dunaliella tertiolecta* and *Monochrysis lutheri*. Patterson (1970) reported an increase in ATP pool size after addition of substrate to a starved activated sludge. Brezonik and Patterson (1971) reported increased ATP per cell upon addition of substrate to activated sludge. The rapid increase in ATP content shown in Figures 27 to 32 indicate an increase in ATP per cell rather than an increase in biomass. These results agree with those of Brezonik and Patterson (1971).

## CHAPTER VIII. CONCLUSIONS AND RECOMMENDATIONS

### Conclusions

It is difficult to evaluate ATP as a biomass parameter since it must be compared to current biomass parameters which, it is believed, are inadequate. There is no absolute biomass parameter to which ATP can be compared. However, working with the available biomass estimates, some conclusions can be made.

Good correlation was observed between ATP and three biomass parameters: chlorophyll *a*, dry weight, and cell number. Cellular ATP did not vary under light and dark conditions, and it was relatively constant under normal or ambient pH conditions. It is proposed, therefore, that ATP could be used as a biomass parameter in studies of aquatic systems. One problem with ATP as a biomass parameter is that it is a measure of total biomass, not just algal biomass. In measuring the ATP content of the unialgal cultures and the lake phytoplankton, no account was taken of the bacterial population. However, in many instances the total biomass is as important or even more important than algal biomass alone. The plankton community consists of phytoplankton, bacteria (free-floating and attached to the phytoplankton), and zooplankton. It seems reasonable that

in pollution or trophic-state studies the biomass of the total community is important, especially considering the symbiotic relationship between bacteria and algae. Too often, the phytoplankton, the bacteria, and the zooplankton have been studied as separate entities when the total community should have been studied.

If a distinction between algal and total biomass is required, then the ATP to chlorophyll ratio can be determined and a rough estimate of the autotrophic nature of a water can be made. Taylor (1967) coined the term "Autotrophic Index" as a measure of the self-feeding or food-producing organisms. The "Autotrophic Index" is defined as the ratio of biomass to chlorophyll. Although the chlorophyll concentration varies considerably under different conditions, it can be used to give a rough estimate of the autotrophic level of a particular lake or stream. Each water, however, because of differences in color, turbidity, and nutritional state, should be monitored to determine the average ambient ATP to chlorophyll ratio. In many waters, the relative abundance of algae compared to bacteria make the distinction between total biomass and algal biomass almost insignificant.

The rapid response of ATP after the addition of toxic substances indicates the feasibility of using ATP in toxicity studies. The data presented in this thesis demonstrates that cellular ATP responds almost immediately. It is

suggested that ATP is an excellent bioassay parameter for toxicity studies.

The response of ATP in nutrient-deficient cultures after addition of nutrients suggests that ATP could be used in limiting nutrient bioassays. However, the qualitative nature of the response must be considered. Like alkaline phosphatase activity, ATP activity is a relative measure of deficiency. Unlike alkaline phosphatase activity which is limited to phosphate-deficient cells, ATP activity can be used for many nutrients. It is suggested that ATP as a limiting nutrient bioassay parameter is an important tool which can be used in a qualitative fashion in nutrient bioassays.

#### Recommendations

The first recommendation of this report is that more theoretical work be performed to determine more about the role and response of ATP under different conditions and in different species. Chemostats should be used in order to maintain algal cultures in a defined state of growth.

Secondly, ATP analysis should be used in routine laboratory and field studies. ATP analysis has a great potential in research and in solving every day problems, but unless it is used in routine studies, it will never be anything but a potentially good tool.



## BIBLIOGRAPHY

- Andrew, J. F. 1968. A mathematical model for the continuous culture of microorganisms utilizing inhibitory substrates. *Biotechn. Bioeng.*, 10: 707-724.
- APHA. 1971. Standard methods for the examination of water and wastewater, 13th ed., New York.
- Arnon, D. I. 1960. The role of light in photosynthesis. *Scientific American*, November.
- Bain, R. C. *et al.* 1968. Nutrients and biological response. San Joaquin master drain - effects on water quality of San Francisco Bay and Delta, Appendix, Part C, Federal Water Pollution Control Administration.
- Balfour, W. M. and F. E. Samson, Jr. 1959. Transphosphorylases in the firefly lantern. *Arch. Biochem. Biophys.*, 84: 140.
- Bauchop, T. and S. R. Elsdon. 1960. The growth of microorganisms in relation to their energy supply. *J. Gen. Microbiol.*, 23: 457-466.
- Bogarad, L. 1962. Chlorophylls, p. 385-404. In R. A. Lewin (ed.) *Physiology and biochemistry of algae*. Academic Press, New York.
- Brezonik, P. L. and J. W. Patterson. 1971. ATP: response to environmental stress in activated sludge. Submitted to *J. Sanit. Eng. Div., Amer. Soc. Civil Eng.*
- Brock, T. D. 1970. *Biology of microorganisms*. Prentice-Hall, Inc., New Jersey. 737 p.
- Brunel, J. *et al.* 1949. The culturing of algae - a symposium. Antioch Press, Ohio.
- Bungay, H. R. III. 1969. Introduction to the symposium on continuous culture, third international fermentation symposium. *Biotechn. Bioeng.*, 11: 773.
- Burlew, J. S. (editor). 1953. *Algal culture from laboratory to pilot plant*. Carnegie Inst. Wash. Publ. # 690, Washington, D.C.

- Carter, B. L. and A. T. Bull. 1969. Studies of fungal growth and intermediary carbon metabolism under steady and non-steady state conditions. *Biotechn. Bioeng.*, 11: 785-804.
- Cole, H. A., J. W. T. Wimpenny *et al.* 1967. The ATP pool in *Escherichia coli*. *Biochem. Biophys. Acta.*, 143: 445.
- Coombs, J. *et al.* 1967a. Chemical composition of *Navicula pelliculosa* during silicon-starvation synchrony. *Plant Physiol.*, 42: 1601-1606.
- Coombs, J. *et al.* 1967b. Changes in concentration of nucleoside triphosphates in silicon-starvation synchrony of *Navicula pelliculosa*. *Exptl. Cell Res.*, 47: 315-328.
- Crietz, G. I. and F. A. Richards. 1955. The estimation and characterization of plankton populations by pigment analysis. *J. Mar. Res.*, 14: 211.
- D'Eustachio, A. J. and E. I. Levin. 1967. Levels of adenosine triphosphate during bacterial growth. *Bact. Proc.*, 67: 121.
- D'Eustachio, A. J. *et al.* 1968. Rapid assay of bacterial populations. *Bact. Proc.*, 21: 13-20.
- Dugdale, R. G. 1967. Nutrient limitation in the sea; dynamics, identification, and significance. *Limnol. Oceanog.* 12: 685-695.
- Droop, M. R. 1966. Vitamin biz and marine ecology. III: an experiment with a chemostat. *J. Mar Biol. Assoc. United Kingdom*, 41: 69-76.
- Elsden, S. R. 1963. The energy requirement for growth of microorganisms. *Bact. Proc.*, 16: XIX.
- Emerson, R. 1929. The relation between maximum rate of photosynthesis and concentration of chlorophyll. *J. Gen. Physiol.*, 12: 609-622.
- Fitzgerald, F. P. and T. C. Nelson. 1966. Extractive and enzymatic analysis for limiting or surplus phosphorus in algae. *J. Phycol.*, 2: 32-37.
- Fitzgerald, F. P. 1968. Detection of limiting or surplus nitrogen in algae and aquatic weeds. *J. Phycol.*, 4: 121-126.

- Fitzgerald, F. P. 1969. Field and laboratory evaluations of bioassays for nitrogen and phosphorus with algae and aquatic weeds. *Limnol. and Oceanog.*, 14: 206-212.
- Fogg, G. E. 1953. The metabolism of algae. Methuen, London.
- Fogg, G. E. 1965. Algal cultures and phytoplankton ecology. Univ. of Wisconsin Press, Wisconsin.
- Forrest, W. W. 1965. Adenosine triphosphate pool during the growth cycle in *Streptococcus faecalis*. *J. Bact.*, 90: 1013.
- Forrest, W. W. and D. J. Walker. 1965. Synthesis of reserve materials for endogenous metabolism in *Streptococcus faecalis*. *J. Bact.*, 89: 1448.
- Franzen, J. S. and S. B. Binkley. 1961. Comparison of the acid soluble nucleotides in *Escherichia coli* at different growth rates. *J. Biol. Chem.*, 236: 515.
- Gerloff, G. C. and F. Skoog. 1954. Cell contents of nitrogen and phosphorus as a measure of their availability for growth of *Microcystis aeruginosa*. *Ecology*, 38: 556-561.
- Gerloff, G. C. and F. Skoog. 1957. Nitrogen as a limiting factor for the growth of *Microcystis aeruginosa* in southern Wisconsin lakes. *Ecology*, 38: 556-61.
- Gerloff, G. C. and P. H. Krombholz. 1966. Tissue analysis as a measure of nutrient availability for the growth of angiosperm aquatic plants. *Limnol and Oceanog.*, 11: 528-537.
- Gerloff, G. C. 1969. Evaluating nutrient supplies for the growth of aquatic plants in natural waters. p. 537-555. In *Eutrophication: causes, consequences, correctives*. Nat'l Academy of Sciences, Washington, D.C.
- Glooschenoko, W. A. and J. E. Moore. 1971. Distribution of chlorophyll *a* and degradation products in Lake Ontario. 34th annual ASLO meeting, Winnipeg, Canada.
- Goldman, C. R. 1960. Primary productivity and limiting factors in three lakes of the Alaska peninsula. *Ecol. Monogr.*, 30: 207-230.
- Goldman, C. R. 1961. Primary productivity and limiting factors in Brooks Lake Alaska. *Verh. Internat. Verein. Limnol.*, 14: 120-124.

- Goldman, C. R. 1962. A method of studying nutrient limiting factors *in situ* in water columns isolated by polyethylene film. *Limnol. Oceanogr.*, 7: 99-101.
- Goldman, C. R. 1963. The measurement of primary productivity and limiting factors in freshwater with carbon-14, p. 103-113. In M. S. Doty (ed.), *Proceedings of the conference on primary productivity measurement, marine and freshwater*, U. S. Atomic Energy Commission, TID-7653, Washington, D.C.
- Goldman, C. R. 1964. Primary productivity and micro nutrient limiting factors in some north american and New Zealand lakes. *Verh. Internat. Verein. Limnol.*, 15: 365-374.
- Goldman, C. R. 1965. Micronutrient limiting factors and their detection in natural phytoplankton populations. In C. R. Goldman (ed.) *Primary productivity in aquatic environments*. Univ. of Calif. Press, California.
- Goldman, C. R. and Carter. 1965. An investigation by rapid carbon-14 bioassay of factors affecting the cultural eutrophication of Lake Tahoe. *J. Water Poll. Control Federation*, 37: 1044-1059.
- Goldman, C. R. 1967. Integration of field and laboratory experiments in productivity studies. In G. H. Lauf (ed.) *Estuaries*, AAAS Publ. No. 83, Washington, D.C.
- Goldman, C. R. *et al.* 1969. Carbon-14 uptake as a sensitive measure of the growth of algal cultures. In *proceedings of the eutrophication biostimulation assessment workshop, June 19-21, 1969*, Univ. of Calif., Berkeley.
- Gylls, F. S. M. 1961. *Biochemist's handbook*. Van Nostrand.
- Hamilton, R. D. and O. Holm-Hansen. 1967. Adenosine triphosphate content of marine bacteria. *Limnol. Oceanogr.*, 12: 319-324.
- Hastings, J. W. 1968. *Bioluminescence*. Annual review of biochemistry, Annual Reviews, Inc., Palo Alto.
- Hee, R. J. 1954. Hemocytometer enumeration of algal cells, an experiment error evaluation, laboratory report of algae research projects. Berkeley Sanit. Eng. Research Lab., Univ. of Calif.

- Herbert, D., R. Ellsworth, and R. C. Telling, 1956. The continuous culture of bacteria: a theoretical and experimental study. *J. Gen. Microb.*, 14: 601-622.
- Holm-Hansen, O. and C. R. Booth. 1966. The measurement of adenosine triphosphate in the ocean and its ecological significance. *Limnol. Oceanog.*, 11: 510.
- Holm-Hansen, O., W. H. Sutcliffe, and J. Sharp. 1968. Measurement of deoxyribonucleic acid in the ocean and its ecological significance. *Limnol. Oceanog.*, 13: 507-514.
- Holm-Hansen, O. 1969. Algae: amounts of DNA and organic carbon in single cells. *Science*, 163: 87-88.
- Holm-Hansen, O. 1970. ATP levels in algal cells as influenced by environmental conditions. *Plant and Cell Physiol.*, 11: 689-700.
- Holm-Hansen, O. 1971. Private communication.
- Hopkins, T. A., H. H. Seliger *et al.* 1967. The excited state products of bioluminescence. *J. Amer. Chem. Soc.*, 89: 7148.
- Jannasch, H. W. and R. F. Vaccaro. 1966. Studies on heterotrophic activity in seawater based on glucose assimilation. *Limnol. and Oceanog.*, 11: 596-607.
- Jenkins, D. and L. Medsker. 1964. Brucine method for determination of nitrate in ocean, estuarine and fresh waters. *Anal. Chem.*, 36: 610-612.
- Joint Industry-Government Task Force on Eutrophication. 1969. Provisional Algal Assay Procedure. Nat'l. Eutrophication Research Program, Environmental Protection Agency, Corvallis, Oregon.
- Kahn, L. and F. T. Brezenski. 1967. Determination of nitrate in estuarine waters. *Envir. Sci. Technol.*, 1: 492-494.
- Keup, L. E. and R. K. Stewart. 1966. Effects of pollution on biota of the Pigeon River, North Carolina and Tennessee, Tech. Advisory and Investigation Activities, FWPCA, USDJ.
- Kono, T. 1968. Kinetics of microbial cell growth. *Biotechn. Bioeng.*, 10: 105-132.

- Kono, T. and T. Asai. 1969. Kinetics of continuous cultivation. *Biotechn. Bioeng.*, 11: 19-36.
- Krauss, R. W. 1962. Inhibitors. In R. A. Lewin (ed.) *Physiology and biochemistry of algae*. Academic Press, New York.
- Lee, C. C. *et al.* 1971. Adenosine Triphosphate content of *Selenastrum capricornutum*. *Applied Microbiology*, 21: 957-958.
- Lehninger, A. L. 1970. *Biochemistry: the molecular basis of cell structure and function*. Worth, New York.
- Lehninger, A. L. 1970. *Bioenergetics*. W. A. Benjamin, Inc., Calif.
- Levine, R. P. 1969. The mechanism of photosynthesis. *Scientific American*, 221: 58-70.
- Lewin, R. A. (editor). 1962. *Physiology and biochemistry of algae*. Academic Press, New York.
- Lyman, G. E. and DeVincenzo, J. P. 1967. Determination of picogram amounts of ATP using the luciferin-luciferase enzyme system. *Anal. Biochem.*, 21: 435.
- Malek, I., K. Beran, and J. Hospodka. 1962. Continuous cultivation of microorganisms. *Proceedings of the second symposium on continuous cultivation of microorganisms*. Academic Press, New York.
- Malek, I. and Z. Fencl. 1966. Theoretical and methodological basis of continuous culture of microorganisms. Academic Press, New York. 655 p.
- Margalef, R. 1967. In *Estuaries*, AAAS Publ. No. 83, Washington, D.C.
- Massux, W. S. and R. F. Jones. 1964. Some interactions of temperature, light intensity, and nutrient concentration during the continuous culture of *Nitzschia closterium* and *Tetraselmis sp.* *Limnol. Oceanog.*, 9: 79-86.
- Matthern, R. O. *et al.* 1969. Effect of total illumination upon continuous *Chlorella* production in a high intensity light system. *Biotechn. Bioeng.*, 11: 863-874.
- McAllister, *et al.* 1964. Marine phytoplankton photosynthesis as a function of light intensity: a comparison of methods. *J. Fish. Res. Bd., Canada*, 21: 1.

- McElroy, W. D. 1947. The energy source for bioluminescence in an isolated system. *Proc. Nat. Acad. Sci.*, 33: 342-350.
- McElroy, W. D. and B. L. Strehler. 1954. Bioluminescence. *Bact. Res.*, 18: 177-184.
- McGauhey, P. H., G. A. Rohlich, E. A. Pearson, M. Tunzi, A. Aleti, and E. J. Middlebrooks. 1968. First progress report, eutrophication of surface water - Lake Tahoe bioassay of nutrient sources. Lake Tahoe Area Council, South Lake Tahoe, Calif., 178 p.
- McGauhey, *et al.* 1969. Second progress report, eutrophication of surface waters - Lake Tahoe laboratory and pilot pond studies. Lake Tahoe Area Council, South Lake Tahoe, Calif., 180 p.
- Menzel, D. W. and J. H. Ryther. 1960. The annual cycle of primary production in the Sargasso Sea off Bermuda. *Deep-Sea Res.*, 6: 351-367.
- Menzel, D. W. and J. H. Ryther. 1961. Nutrients limiting the production of phytoplankton in the Sargasso Sea. *Deep-Sea Res.*, 7: 276-281.
- Menzel, D. W. and J. H. Ryther. 1964. The composition of particulate organic matter in the western North Atlantic. *Limnol. Oceanogr.*, 9: 179-186.
- Menzel, D. W. and J. J. Goering. 1966. The distribution of organic detritus in the ocean. *Limnol. Oceanogr.*, 11: 333-337.
- Middlebrooks, E. J. *et al.* 1969. Proceedings of the eutrophication biostimulation assessment workshop. Berkeley, Calif.
- Middlebrooks, E. J. *et al.* 1971. Biostimulation and algal growth kinetics of wastewater. *J. Water Pollution Control Federation*, 43: 454-473.
- Monod, J. 1949. The growth of bacterial culture. *Ann. Rev. Microbiology*, III.
- Murphy, J. and J. P. Riley. 1962. A modified single solution method for the determination of phosphate in natural waters. *Analyt. Chim. Acta.*, 27: 31-36.

- Murray, S. *et al.* 1971. Evaluation of algal assay procedures - PAAP batch test. J. Water Pollution Control Federation, 43: 1991-2003.
- Myers, J. 1946. Culture conditions and the development of the photosynthetic mechanism. III. Influence of light intensity on cellular characteristics of *Chlorella*. J. Gen. Physiol., 29: 419-427.
- Myers, J. 1962. Laboratory cultures. In R. A. Lewin (ed.) Physiology and biochemistry of algae. p. 603-616. Academic Press, New York.
- Novick, A. and L. Szilard. 1950. Experiments with the chemostat on spontaneous mutation of bacteria. Proc. Nat'l. Acad. Sci., U.S., 36: 708.
- Odum, H. T. *et al.* 1958. The chlorophyll *a* of communities. Publ. Inst. Mar. Sci., Univ. Texas, 5: 65-97.
- Odum, H. T. and S. Nixon. 1970. A model for photoregeneration in brines. In ESE notes, Univ. of North Carolina.
- Odum, H. T. *et al.* 1971. Adaptations for photoregeneration cycling. In Press.
- Oswald, W. J. and C. G. Golueke. 1966. Eutrophication trends in the United States - a problem? 38(6): 964-975.
- Oswald, W. J. 1967. The meaning, the measurement, and the manipulation of algal growth potential. Presented before the New York State Health Department, Albany, New York.
- Oswald, W. J. and S. A. Gaonkar. 1969. Batch assays for determination of algal growth potential. In E. J. Middlebrooks (ed.) Proceedings of the eutrophication biostimulation assessment workshop. Berkeley, Calif.
- Parsons, T. R. and T. D. H. Strickland. 1962. Oceanic detritus. Science, 136: 313-314.
- Patterson, J. W. and P. L. Brezonik. 1969. Toxicity measurements in activated sludge-discussion. Jour. Sanit. Eng. Div., Amer. Soc. Civ. Eng., 95: 775-780.
- Patterson, J. W. 1970. The application of biochemical activity indices to an activated sludge system. Doctoral Dissertation, Univ. of Florida, Gainesville.



- Patterson, J. W. *et al.* 1969. Sludge activity parameters and their application to toxicity measurements in activated sludge. Purdue Industrial Waste Conf. Proceedings, May, 1969, Lafayette, Indiana.
- Pearson, E. A., E. J. Middlebrooks, M. Tunzi, A. Aletti, P. H. McGauhey, and G. A. Rohlich. 1969. Kinetic assessment of algal growth. 64th National Meeting American Institute of Chemical Engineers, 16-20 March 1969, New Orleans, Louisiana.
- Pipes, W. O. 1962. Carbon dioxide-limited growth of *Chlorella* in continuous cultures. *Appl. Microbiol.*, 10: 281-288.
- Pipes, W. O. and S. P. Kontsoyannis. 1962. Light-limited growth of *Chlorella* in continuous cultures. *Appl. Microbiol.*, 10: 5.
- Provasoli, L., S. H. Hunter, and I. J. Pinter. 1951. Destruction of chloroplasts by streptomycin. Cold Spring Harbor Symposium, *Quant. Biol.*, 16: 113-120.
- Provasoli, L., I. Pinter, and L. Packer. 1951. Use of antibiotics in obtaining pure cultures of algae. *Proc. Am. Soc. Protozool.*, 2: 6.
- Putnam, H. D. 1966. Limiting factors for primary production in a west coast Florida estuary. Third International Conference on Water Pollution Research, Munich, Germany.
- Rabinowitch, E. I. and Govindjee. 1965. The role of chlorophyll in photosynthesis. *Scientific American*, 213: 74-83.
- Rhodes, W. C. and W. D. McElroy. 1958. The synthesis and function of luciferyl-adenylate and oxyluciferyl-adenylate. *J. Biol. Chem.*, 233: 1528.
- Richards, F. A. and T. G. Thompson. 1952. The estimation and characterization of plankton populations by pigment analysis. II a spectrophotometric method for the estimation of plankton pigments. *J. Mar. Res.*, 11: 156.
- Round, F. E. 1965. Light and temperature: some aspects of their influences on algae. In D. F. Jackson (ed.) *Algae, man, and the environment*. Syracuse University Press, New York.

- Ryther, J. H. and C. S. Yentsch. 1957. The estimation of phytoplankton production in the ocean from chlorophyll and light data. *Limnol. Oceanog.*, 2: 281-286.
- Ryther, J. H. and D. W. Menzel. 1964. On the production, composition, and distribution of organic matter in the Western Arabian Sea. *Deep-Sea Research*, 12: 199-209.
- Santarius, K. A. and U. W. Heber. 1965. Loss of ATP synthesis caused by freezing and its relationship to frost hardiness problems. *Plant Physiol.*, 39: 712-719.
- Sargent, M. C. 1940. Effect of light intensity on the development of the photosynthetic mechanism. *Plant. Physiol.*, 15: 275-290.
- Seliger, H. H. and W. D. McElroy. 1959. Quantum yield on the oxidation of firefly luciferin. *Biochem. Biophys. Res. Com.*, 1: 21.
- Seliger, H. H. and W. D. McElroy. 1960. Spectral emission and quantum yield of firefly bioluminescence. *Arch. Biochem. Biophys.*, 88: 136.
- Shannon, E. E. 1970. Eutrophication-trophic state relationships in north and central Florida lakes. Doctoral Dissertation, Univ. of Florida, Gainesville.
- Shapiro, J. and R. Ribeiro. 1965. Algal growth and sewage effluent in the Potomac Estuary. *J. Water Poll. Control Federation*, 37: 1034-1043.
- Slezak, J. and B. Sikyta. 1967. Growth of *Escherichia coli* B in a continuous culture under limitation by inorganic phosphate. *Folia. Microb.*, 12: 441-446.
- Spencer, C. P. 1952. On the use of antibiotics for isolating bacteria-free cultures of marine phytoplankton organisms. *J. Mar. Biol. Assoc. United Kingdom*, 31: 97-106.
- Spoehr, H. A. and H. W. Milner. 1949. The chemical composition of *Chlorella*; effect of environmental conditions. *Plant Physiol.*, 24: 120-149.
- Steeemann-Nielsen, E. 1951. Measurement of the production of organic matter in the sea by means of carbon-14. *Nature*, 1967: 684-685.

- Strickland, J. D. H. and T. R. Parsons. 1968. A practical handbook of seawater analysis. Bulletin 167 of the Fisheries Research Board of Canada, Ottawa.
- Syrett, P. J. 1958. Respiration rate and internal ATP concentration in *Chlorella*. Arch. Biochem. Biophysics, 75: 117.
- Syrett, R. J. 1962. Nitrogen assimilation, p. 171-188. In R. S. Lewin (ed.) Physiology and biochemistry of algae. Academic Press, New York.
- Taylor, M. P. 1967. Thermal effects on the periphyton community in the Green River. Tenn. Valley Authority, Div. of Health and Safety, Chattonooga, Tenn.
- Tempest, D. W. *et al.* 1965. Magnesium-limited growth of aerobacter aerogenes in a chemostat. J. Gen. Micro., 39: 355-366.
- Tempest, D. W. *et al.* 1967. Magnesium-limited growth of bacillus subtilis in pure and mixed cultures in a chemostat. J. Gen. Microb., 49: 139-147.
- Tempest, D. W. and D. C. Ellwood. 1969. The influence of growth conditions on the composition of some cell wall components of Aerobacter aerogenes. Biotechn. Bioeng., 11: 775-784.
- Vaccaro, Ralph F. 1969. The response of natural microbial populations in seawater to organic enrichment. Limnol. Oceanogr., 14: 726.
- Vollenweider, R. A. 1970. Scientific fundamentals of the eutrophication of lakes and flowing waters, with particular reference to nitrogen and phosphorus as factors in eutrophication. O.E.D.C. Report, Paris, France.
- Weber, C. I. and B. H. McFarland. 1969. Periphyton biomass-chlorophyll ratio as an index of water quality. Presented at 17th Annual Meeting of Midwest Benthological Society, Kentucky.
- Weddle, C. L. and D. Jenkins. 1969. The viability and activity of activated sludge. 42nd Annual WPCF Conf., Dallas, Texas.
- Wetherell, D. F. 1958. Obligate phototrophy in *Chlamydomonas euganetos*. Physiol. Plantarum, 11: 260-274.

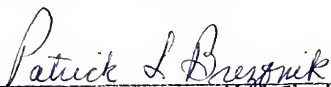
- Wimpenny, J. W. T. 1967. Adenosine triphosphate pool levels and turnover in microorganisms. *Biochem. J.*, 102: 54.
- Yano, T. and S. Koga. 1969. Dynamic behavior of the chemostat subjected to substrate inhibition. *Biotechn. Bioeng.*, 11: 139-154.
- Yentsch, C. S. and J. H. Ryther. 1957. Short term variations in phytoplankton chlorophyll and their significance. *Limnol. Oceanog.*, 2: 140-142.
- Yentsch, C. S. and D. W. Menzel. 1965. A method for determination of phytoplankton chlorophyll and phaeophytin by fluorescence. *Deep-Sea Res.*, 10: 221-231.
- Yentsch, C. S., and R. H. Backus, and A. King. 1964. Factors affecting the vertical distribution of bioluminescence in the eutrophic zone. *Limnol. Oceanog.*, 9: 519-524.

## BIOGRAPHICAL SKETCH

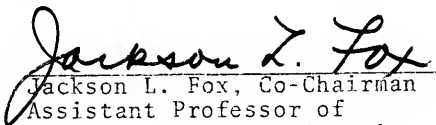
Francis Xavier Browne was born October 11, 1943, at Brooklyn, New York. In June, 1961, he was graduated from Xaverian High School. In June, 1966, he received the degree Bachelor of Civil Engineering from Manhattan College. In September, 1966, he enrolled in the Graduate School of Manhattan College. He worked as a research assistant in the Department of Civil Engineering until January, 1967, when he received a Graduate Fellowship. In January, 1968, he received the degree of Master of Civil Engineering with a major in Sanitary Engineering. From 1967 to 1968 he was employed as a consulting sanitary engineer at Hydrotechnic Corporation, New York. From September, 1968, until the present time he has been enrolled in the Graduate School of the University of Florida while he has pursued his work toward the degree of Doctor of Philosophy.

Francis Xavier Browne is married to the former June Marie Kendall. He is a member of the Water Pollution Control Federation, and the American Society of Limnology and Oceanography.

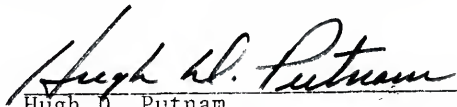
I certify that I have read this study and that in my opinion it conforms to acceptable standards of scholarly presentation and is fully adequate, in scope and quality, as a dissertation for the degree of Doctor of Philosophy.

  
Patrick L. Brezonik, Chairman  
Associate Professor of  
Environmental Engineering

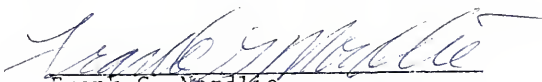
I certify that I have read this study and that in my opinion it conforms to acceptable standards of scholarly presentation and is fully adequate, in scope and quality, as a dissertation for the degree of Doctor of Philosophy.

  
Jackson L. Fox, Co-Chairman  
Assistant Professor of  
Environmental Engineering

I certify that I have read this study and that in my opinion it conforms to acceptable standards of scholarly presentation and is fully adequate, in scope and quality, as a dissertation for the degree of Doctor of Philosophy.

  
Hugh D. Putnam  
Professor of Environmental  
Engineering

I certify that I have read this study and that in my opinion it conforms to acceptable standards of scholarly presentation and is fully adequate, in scope and quality, as a dissertation for the degree of Doctor of Philosophy.

  
Frank G. Nördlie  
Associate Professor of Zoology

This dissertation was submitted to the Dean of the College of Engineering and to the Graduate Council, and was accepted as partial fulfillment of the requirements for the degree of Doctor of Philosophy.

December, 1971

  
Dean, College of Engineering

\_\_\_\_\_  
Dean, Graduate School



UNIVERSITY OF FLORIDA



3 1262 08666 936 2